



Greenery as a mitigation and adaptation strategy to urban heat

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Abstract | The absence of vegetation in urban areas contributes to the establishment of the urban heat island, markedly increasing thermal stress for residents, driving morbidity and mortality. Mitigation strategies are, therefore, needed to reduce urban heat, particularly against a background of urbanization, anthropogenic warming and increasing frequency and intensity of heatwaves. In this Review, we evaluate the potential of green infrastructure as a mitigation strategy, focusing on greenery on the ground (parks) and greenery on buildings (green roofs and green walls). Green infrastructure acts to cool the urban environment through shade provision and evapotranspiration. Typically, greenery on the ground reduces peak surface temperature by 2–9 °C, while green roofs and green walls reduce surface temperature by ~17 °C, also providing added thermal insulation for the building envelope. However, the cooling potential varies markedly, depending on the scale of interest (city or building level), greenery extent (park shape and size), plant selection and plant placement. Urban planners must, therefore, optimize design to maximize mitigation benefits, for example, by interspersing parks throughout a city, allocating more trees than lawn space and using multiple strategies in areas where most cooling is required. To do so, improved translation of scientific understanding to practical design guidelines is needed.

Sensible heat

Heat transfer that results in a change in temperature between objects, without changing the volume or pressure.

The urban heat island (UHI) effect describes an increase in temperature of dense urban areas compared with their suburban or rural surroundings^{1,2} (FIG. 1a). The UHI arises through shifts in energy fluxes associated with land use change — specifically, an increase in solar absorption, sensible heat and heat trapping³, and a corresponding reduction in evapotranspiration — as well as increased anthropogenic heat from buildings and vehicles⁴ (FIG. 1b). UHI intensity typically varies between 0.4 °C and 11 °C (REF.⁵), and is more pronounced at night⁶, exposing residents to higher thermal stress.

Such exposure can have adverse impacts for human health, producing increased mortality and morbidity, especially amongst low-income and vulnerable populations, such as the elderly^{7,8}. These impacts are heightened during heatwave events when temperatures are already amplified^{9,10}. For example, during the 2003 European heatwave, it is estimated that the UHI contributed to 50% of the total deaths in the West Midlands, UK¹¹. The same heatwave also caused an estimated 70,000 excess deaths across Europe¹², most prominent in urban locations such as Paris, France¹³. Analyses in various other global cities also indicate greater mortality in urban regions during heatwaves owing to the UHI effect, including Shanghai¹⁴, Hong Kong¹⁵, Ho Chi Minh¹⁶, Athens¹⁷ and London¹⁸.

Accordingly, there is an urgent need to mitigate urban warming and its deleterious impacts, especially against a background of increased urbanization and anthropogenic warming. Urbanization, for example, is widely regarded to increase local temperatures, and, thus, UHI intensity, in the future^{19,20}. While the impacts of climate change are more variable — for instance, increasing UHI intensities in Chicago²¹, Beijing²² and Melbourne²³, but decreasing them in Paris²⁴ and Brussels²⁵ — the combined impact of both factors is anticipated to exacerbate the UHI effect^{26,27} and, thereby, UHI-related mortality^{11,28}. Moreover, given that heatwaves interact non-linearly with UHIs to amplify urban heat stress in the present climate²⁹, projected increases in heatwave frequency and intensity might similarly magnify heat stress further in the future^{30,31}.

Government sectors and policymakers have, thus, considered several active and passive strategies to address the UHI effect. These methods include: reducing shortwave and longwave absorption by modifying the reflectance properties of urban surfaces — increasing the albedo of building materials and surfaces (particularly roofs)^{32,33}; designing urban geometry to minimize heat gain and facilitate the release of stored heat and dissipation via urban ventilation^{34,35}; lessening anthropogenic heat creation by increasing energy efficiency and reducing vehicle use³⁶; expanding blue spaces such as lakes

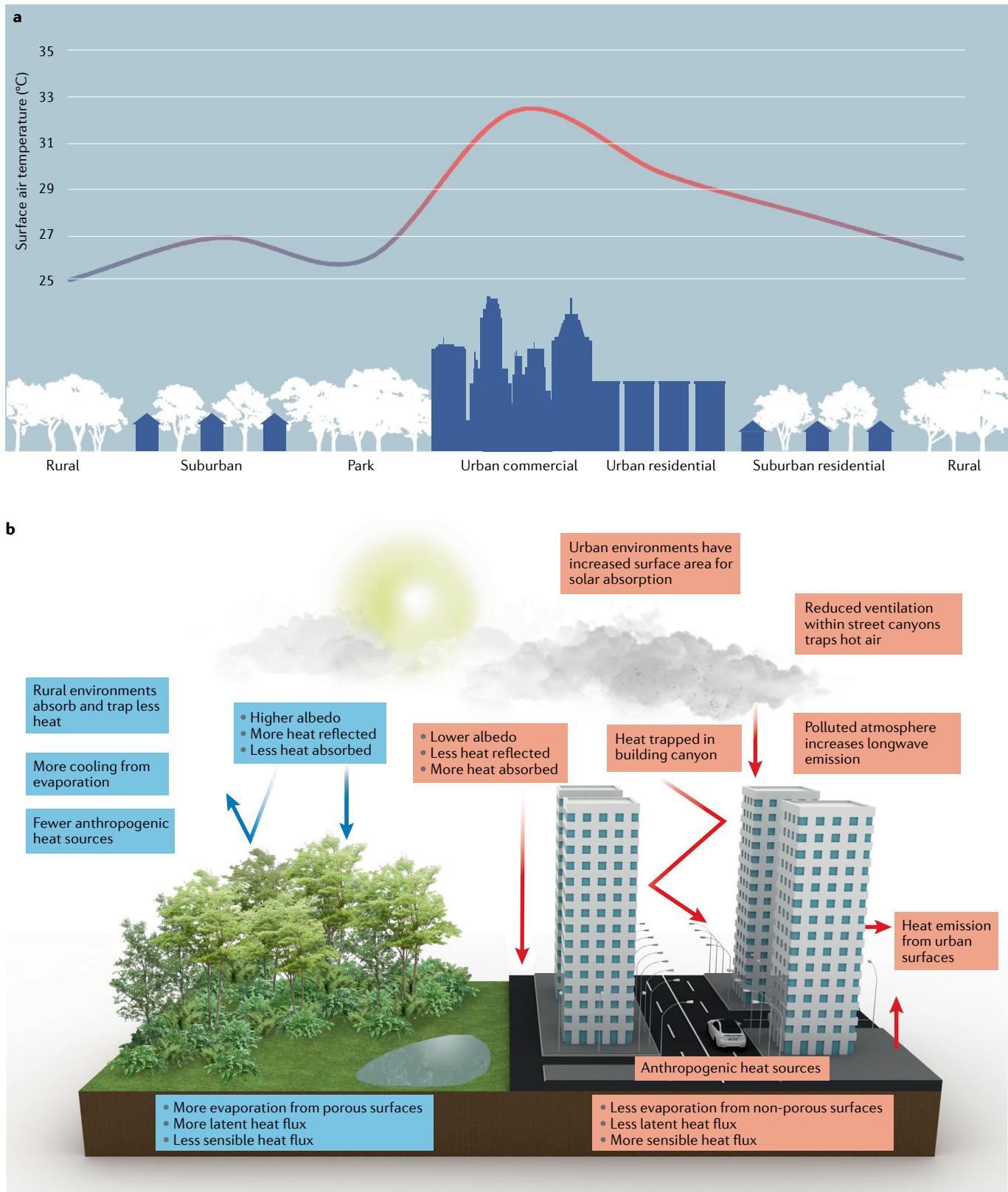
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and rivers³⁷; and promoting cooling via evapotranspiration with urban greening. Indeed, green infrastructure — encompassing green roofs, green facades and parkland expansion — has been regarded as an effective mitigation strategy for urban heat^{38,39}, and, at the same time,

brings ecosystem services and co-benefits in terms of carbon sequestration⁴⁰, phytoremediation⁴¹, improved air quality⁴² and promoting biodiversity^{43,44}.

In this Review, we examine urban greeneries as a mechanism to mitigate urban heat. We begin by outlining the

◀ Fig. 1 | **The urban heat island effect.** **a** | A typical urban heat island profile, showing higher air temperature in built-up areas and lower temperature in rural areas with more greenery coverage. **b** | Factors contributing to the urban heat island effect, highlighting significant changes in heat and air movement when rural land is urbanized. Red boxes indicate warming mechanisms and blue boxes indicate cooling mechanisms.

physical mechanisms through which greenery contributes to cooling, followed by a discussion of the different forms of greenery — green parks, green roofs and green walls — and their cooling potential. The implications for planning and design of buildings and cities are subsequently considered, followed by the future needs and priorities of the research community.

Mechanisms of greenery-related cooling

Regardless of the specific approach adopted, green infrastructure acts to cool urban environments through various mechanisms: evapotranspiration⁴⁵, shade provision⁴⁶ and increased albedo^{32,47}, the combination of which reduces the thermal load on the built environment and its inhabitants (FIG. 2).

Heat flux interception from the plant canopy (and, thus, provision of shade) is one of the most direct and effective means of cooling the urban microclimate^{48,49}, dominating the cooling potential of green infrastructure. Depending on the density of their canopies, plants are able to intercept 70–90% of incoming solar radiation^{50,51}, still reaching 50% for deciduous trees during winter when leaf counts are substantially lowered⁵². This reduction in both shortwave and longwave radiation substantially cools urban surfaces such as buildings, roads and pavements, in turn, reducing the mean air temperature of surroundings. Shade from green infrastructure can also lower energy requirements for cooling, reducing anthropogenic heat sources and potentially reducing energy savings by 20–80%^{53–55}.

Vegetation further allows evapotranspiration. Evapotranspiration uses solar energy to convert liquid water into water vapour, thereby, replacing sensible heat with latent heat^{56–58}. Thus, compared with impervious urban environments where sensible heat gain occurs owing to an absence of water, this evaporative cooling effect provides the important function of lowering the Bowen ratio and temperature of surrounding landscapes⁵⁹. The reduction in sensible heat gain also acts to lower plant canopy surface temperature and decrease longwave emission to surroundings⁴⁵.

The presence of greenery can further enhance the albedo of highly urbanized environments. For example, the albedo of built-up areas varies from ~0.1 to 0.2 (REFS^{50,51}), whereas the albedo of plants can reach close to 0.3 (REF.⁶²). Raising the albedo increases the proportion of incoming radiation that is reflected, thus, decreasing the component that is absorbed and, therefore, able to increase surface temperatures³². Given the limited range of albedo values for plants⁶³, however, the cooling potential arising from albedo changes is lower than that of shading provision and evaporative cooling.

Greenery on ground

Retaining or reintroducing green spaces such as gardens or parks offers one such strategy to mitigate the UHI effect. As discussed, shade provision from vegetation

canopies blocks shortwave and longwave radiation, while also promoting evapotranspiration, lowering longwave emission and, in the case of large urban parks, minimizing anthropogenic heat sources. Through a combination of these factors, green spaces provide an effective means to lower UHI intensity, as revealed through numerous field measurements^{64–69}, numerical simulations^{70–72} and remote sensing^{73–76}.

However, while almost all studies reveal temperature reductions owing to the presence of greenery, the magnitude of cooling varies substantially. For instance, a meta-analysis of 24 studies covering tropical and temperate climates indicates air temperature cooling of 0.94°C (REF.³⁸), whereas another based on 89 studies suggests cooling of 1.5–3.5 °C (REF.⁷⁷). Analyses focusing on one city further show even larger temperature reductions, reaching 4.52 °C in Changchun, China⁷⁸, and 6.82 °C in Nagoya, Japan⁷⁶. Indeed, compiling 30 published studies spanning diverse geographic regions indicates that green parks act to cool air temperature by an average of ~3 °C, with a range of 2–4 °C (FIG. 3).

When assessing surface temperature (rather than air temperature), the cooling potential of green parks is larger (FIG. 3a), mainly due to better thermal conductivity of solid surfaces compared with air. Remote sensing technologies — which provide estimates of cooling over relatively large (60–120 m) spatial footprints — reveal average surface temperature reductions of 4.2 °C, with a range of 1.9–6.7 °C. By contrast, on-site measurements — which are able to capture temperature changes with higher granularity — document average reductions of 14 °C, with a range of 9.2–19 °C (FIG. 3a).

Thus, while the cooling potential of green parks is clear, so too are the contrasts in quantitative estimates. This variability is linked to the methodologies adopted for measurement (on-site measurements at pedestrian height as compared with remote sensing techniques, which are averaged values over large areas), as well as differences in climate, the size and shape of the park, and plant selection and placement (FIG. 4), as will now be discussed.

Climate

The cooling potential of urban greenery on the ground is influenced by both diurnal^{79,80} and seasonal cycles^{67,78}. On diurnal timescales, maximum temperature reductions associated with greening tend to occur during the day. In Hong Kong, for example, an urban park was found to be 8 °C cooler than its surroundings in the day, whilst only 2 °C cooler at night⁸¹. This temporal difference in cooling potential can be attributed to the contrasting diurnal temperature gradients; during the day, exposure to direct solar radiation produces large differences between green and urban spaces, but, by nightfall, heat is emitted back to the atmosphere in urban environments as longwave radiation, minimizing urban–rural contrasts.

Variability in cooling potential is also strongly evident on seasonal timescales, reaching peak amplitude during the summer. For instance, in Nagoya, Japan, park-related cooling was much larger in summer (1.9 °C) compared with winter (0.3 °C)⁸² (FIG. 4a). These differences arise

Evapotranspiration

The combined processes of evaporation of water from the soil, as well as plant transpiration, where water is transported from the soil through the roots and exits via the leaf stomata and into the atmosphere as water vapour.

UHI intensity

The temperature difference between urban and rural areas; either surface or air temperature can be used.

Albedo

The ratio of reflected radiation over total incident radiation on a surface, indicating its overall reflecting potential. Albedo values can range from 0 to 1, with 1 meaning all radiation is reflected and 0 indicating that all radiation is being absorbed.

Latent heat

Heat transfer that results in a change in state (such as liquid into vapour), without changing the temperature.

Bowen ratio

The ratio of sensible heat flux to latent heat flux above a surface that contains moisture. Commonly used in meteorological and hydrological studies, it is an indication of the abundance of water over surfaces, as the presence of moisture will directly influence latent heat flux density.

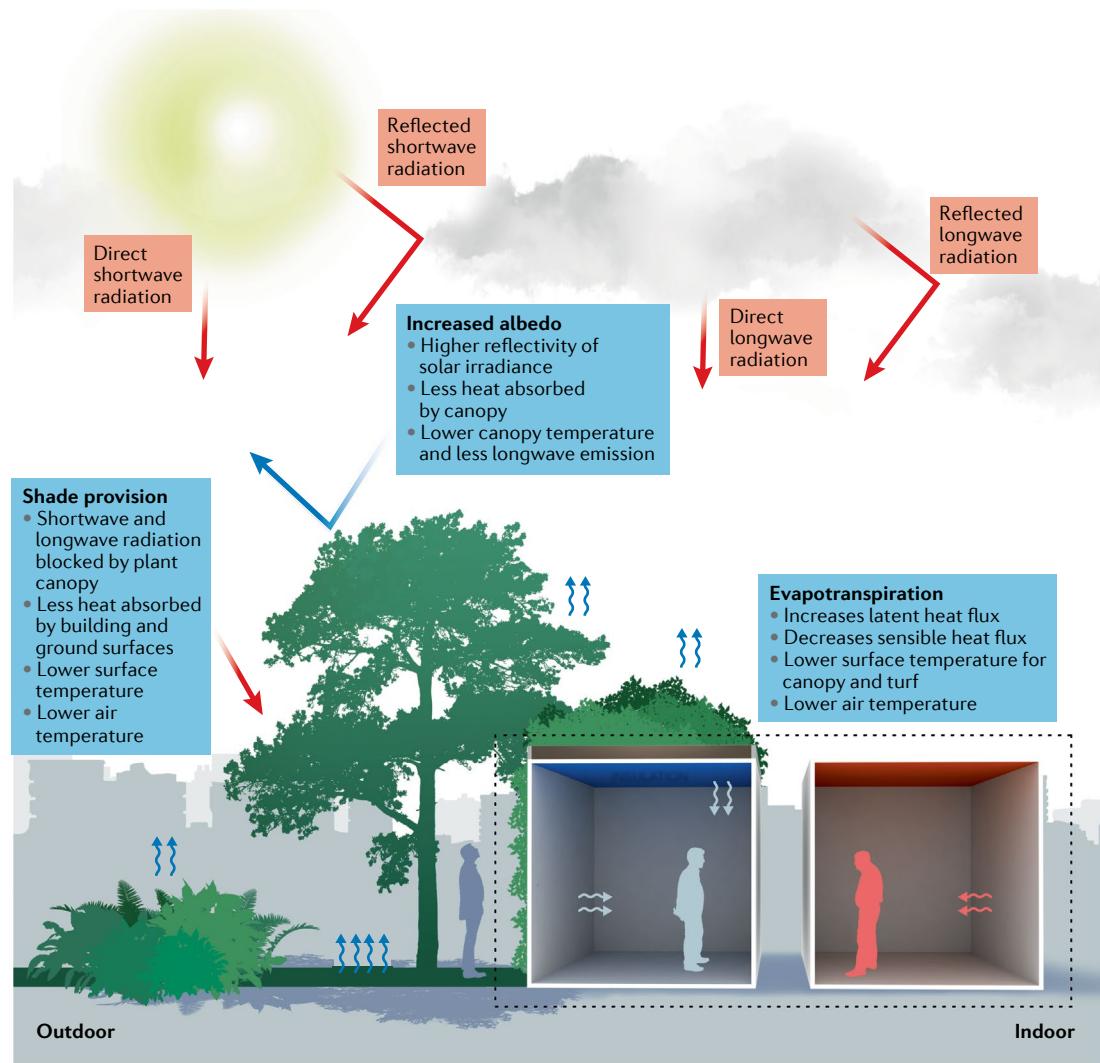


Fig. 2 | Greenery-related cooling mechanisms in the urban environment. Urban greenery acts to modify shade provision, evapotranspiration and albedo. The combination of these three mechanisms reduces sensible heat gain, thereby, lowering heat gain and surface temperatures. Red boxes indicate warming mechanisms and blue boxes indicate cooling mechanisms.

through the strong dependence on shade potential and evapotranspiration, which is drastically reduced in autumn and winter when trees shed their leaves and canopy cover is reduced.

While the magnitude of temperature reduction differs between seasons, the temperature reduction differential has no apparent correlation to specific climatic regions^{38,77}. Indeed, greenery on the ground has been shown to be effective in providing cooling in tropical⁸³, subtropical⁸⁴ and temperate⁸⁵ climates, as well as humid⁸⁶ and arid⁸⁷ regions. Given the small interquartile range of air temperature reductions across studies (FIG. 3a), greenery on the ground is, therefore, an effective mitigation strategy for urban heat, regardless of the locale.

Size and shape of park

In addition to climate, the size and shape of the green space also exerts a strong influence on the cooling potential of urban parks^{88–91}. Larger parks tend to have a more pronounced cooling effect, owing to the net decrease in sensible heat flux and reduced anthropogenic heat

sources. For instance, in Suzhou, China, the average cooling effect increases from 1.75 °C for small (<4 ha), 2.66 °C for medium (4–10 ha) to 3.32 °C for large green spaces (>10 ha)⁹². The size at which peak cooling occurs, however, exhibits pronounced variability, owing to the background climate and urban context (FIG. 3a). In Fuzhou, China, for example, the most efficient cooling occurs at ~4.5 ha (REF. 93), whereas it is 3 ha in Taipei, Taiwan⁹⁴, 5.6 ha in Leipzig, Germany⁹⁵, and 14 ha in Chongqing, China⁹⁶. The threshold value of efficiency (TVoE) for park size (FIG. 4c) has provided some indicative values of park size to which temperature reduction potential due to park size starts to plateau⁹⁷. The TVoE size can range from 0.5 ha to 0.69 ha in temperate cities such as Copenhagen⁹⁸ and Rome⁹⁹, to 0.6 ha to 0.95 ha in tropical cities in Asia¹⁰⁰.

The cooling potential of smaller green spaces is often contradictory. Some research indicates that small green areas have the potential to provide air temperature reductions comparable to large parks^{67,77}. However, urban geometry and prevailing wind conditions become

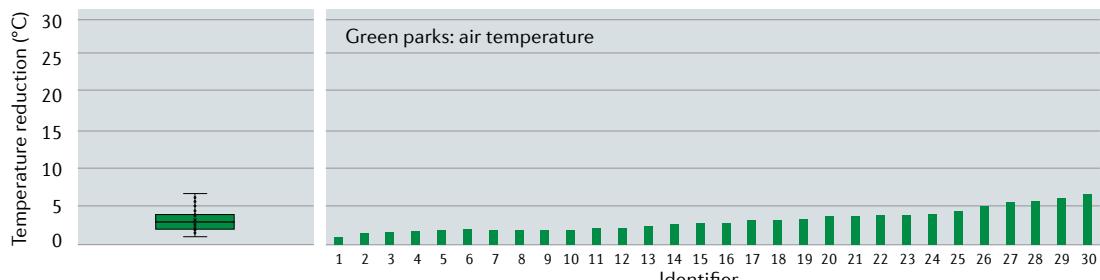
Threshold value of efficiency (TVoE). The value to which an increase in green space ceases to provide substantial cooling.

important factors, and if such conditions are not favourable, small areas can become warmer, not cooler^{76,94}. Indeed, such small spaces are often more susceptible to urban and anthropogenic influences, increasing sensible heat gain, for example, through a greater percentage of paved surfaces. Accordingly, 14 of 61 small parks in Taipei were, on average, 0.42 °C hotter than their surroundings⁹⁴, and a small 1.5-ha inner-city park in Melbourne also experienced a 0.2 °C increase in air temperature during the early part of the day¹⁰¹. The cooling effect beyond park boundaries can also decrease with size. For large parks of more than 100 ha, the cooling effect might extend a few hundred metres beyond the park periphery, whereas the cooling effect of small green

spaces (less than 0.1 ha) may not even extend beyond their boundaries (FIG. 4c).

Similar to size, park shape also influences the cooling effect¹⁰². Regularly shaped parks such as square or circular spaces have been found to exhibit higher cooling efficiency, which drops as the shape gets more complex⁷⁵ (FIG. 4d). In addition, the cooling efficiency of parks is maximized when green spaces are polygonal (circular or regular polygons) compared with linear (long and narrow)^{103,104}. This difference can partly be attributed to the influence of park shape on plant selection. Linear parks tend to consist of identical tree species and often lack smaller trees and shrubbery. Accordingly, they are prone to heat invasion from areas

a Greenery on the ground



b Greenery on buildings

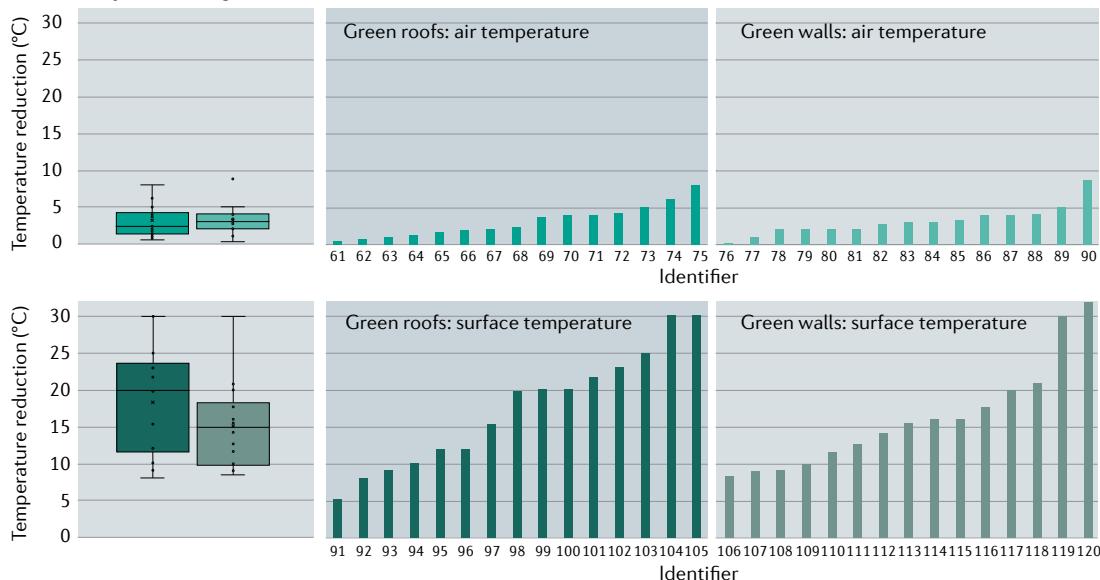
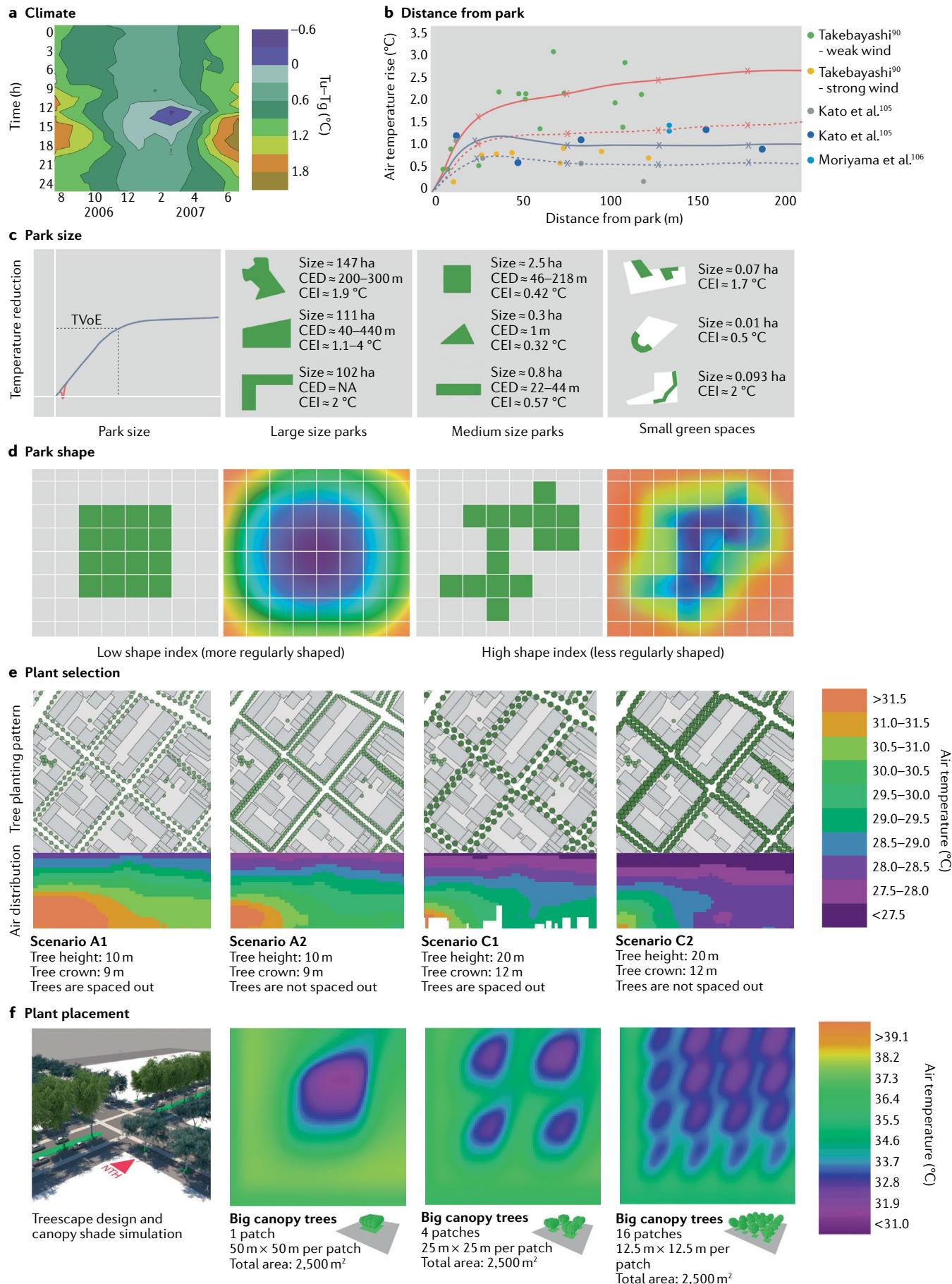


Fig. 3 | Average greenery-related peak temperature reductions. Maximum recorded air and surface temperature reductions associated with greenery on the ground (panel a) and greenery on buildings (panel b) from previously published studies (see Supplementary Table 1 for identifier information). Greenery on the ground are separated by studies assessing temperature changes through remote sensing or field measurements and greenery on buildings are grouped by green roofs and green walls. Box and whisker plots on the left are produced from the individual studies shown on the right.



◀ Fig. 4 | Factors contributing to temperature reduction for ground-level greenery. **a** | Climate: air temperature difference between urban (T_u) and green (T_g) spaces in Nagoya, central Japan, revealing greater cooling potential in the summer. **b** | Distance from park: in situ (dotted points) and modelled (lines) air temperature rise with distance from the park in several areas of Japan, revealing cooling within 50 m from the green space. Orange lines indicate results from an isotropic diffusion model and blue lines indicate results from an incorporating buoyancy model. Solid and dashed lines indicate model simulations under weak and strong wind conditions, respectively. **c** | Park size: the concept of threshold value of efficiency (T_{VoE}) for park cooling based on size, with indication of possible negative cooling for small parks (red line, left panel). Cooling effect distance (CED) and cooling effect intensity (CEI) for large, medium and small green parks of different shapes, revealing that CED diminishes when park size decreases. **d** | Park shape: illustration of air temperature associated with parks of contrasting shape index, indicating that regularly shaped parks cool a larger area. **e** | Plant selection: cross-sectional air temperatures associated with different model scenarios of tree height, canopy size and planting density in Montreal, revealing greater cooling potential as all three factors increase. **f** | Plant placement: visualization of tree canopy shade and ENVI-met simulation showing the difference in air temperature owing to different tree arrangements, revealing greater cooling potential for multiple, smaller tree patches. Panel **a** is adapted with permission from REF⁸². Panel **b** is adapted with permission from REF⁹⁰. Panel **c** is adapted with permission from REF¹⁸⁶. Panel **e** is adapted with permission from REF¹⁸⁷. Panel **f** is adapted with permission from REFS^{154,108}.

outside the green spaces and, thus, display lesser overall temperature reduction. Polygonal green spaces, by contrast, tend to trap the cooled air more efficiently via small trees and shrubs, thereby, maintaining a larger temperature differential¹⁰³.

Cooling effect outside parks

Although green spaces are cooler than the built environment, they only take up a small portion of the entire urban landscape. Therefore, it is important for the cooler air generated by green spaces to be able to permeate and cool surrounding parts of the built environment, so that cooling benefits are tangible over a larger area, mitigating the UHI. Much effort has, thus, been put into understanding how these spaces can influence their surrounding environments^{90,105,106}.

How far greening-related temperature reductions permeate varies markedly. In Beijing, China, for example, the cooling effects of 30 urban parks ranges from 2.3 °C to 4.8 °C and extends 35–840 m outside the park, the distance being governed by park size and characteristics of the surrounding environment⁷³. An area with high building density, for instance, can impede air movement and hinder the exchange of cooled air from parks to their surroundings. In London, UK, park-related cooling was apparent up to 330 m away from the green space, with the distance of cooling again scaled linearly with green space area, but also tree canopy extent¹⁰⁷. Cooling is further evident up to 1.1 km for a 156-ha park in Gothenburg, Sweden⁷⁹. Overall, cooling potential is increased under stronger wind conditions, suggesting that the surrounding estate should be designed to maximize ventilation (FIG. 4b).

Thus, although the spatial extent of cooling varies considerably, it is clear that green spaces have the ability to provide cooling not just within the park (FIG. 4c,d) but also outside its boundaries (FIG. 4b), particularly when cooling potential is maximized through arranging parks at a minimum size (at least 1 ha) and placing at appropriate intervals in the urban area (less than 1 km).

Leaf area index (LAI). Total one-sided leaf area per unit horizontal ground surface.

Plant selection and placement

Size alone does not guarantee maximum temperature reduction, with plant selection and placement also playing a fundamental role in explaining heterogeneous park-related cooling (FIGS 3a,4e,f). For example, owing to their larger canopy (and, hence, shade) and evapotranspiration characteristics, trees provide greater cooling potential in comparison with shrubs and lawns^{108,109}. In particular, in situ estimates from Germany indicate mean radiant temperature reductions of 39.1 °C under trees but only 7.5 °C on grassland¹¹⁰.

The cooling potential of trees themselves, however, varies markedly, owing to contrasting plant functional traits, including canopy size and leaf area, both of which influence shade provision^{111–113}. Canopy coverage, as estimated by the leaf area index (LAI), for example, is positively correlated to temperature reduction¹¹⁴, the larger the canopy, the larger the cooling (FIG. 4e). Tree species such as *Caesalpinia pluviosa*, with dense canopy and large coverage, can provide more than 90% solar attenuation and are ideal for improving the urban microclimate¹¹⁵. Dense canopies have also been shown to maximize cooling¹¹⁶, reducing radiant exposure at ground level by up to 92% (REF¹¹⁷). Similar to shade provision, plant evapotranspiration is also species-specific¹¹⁸, with maximum latent heat flux varying by up to ~760 W m⁻² (REF¹¹⁹).

Interactions between vegetation and buildings must also be considered, especially with small parks, given their influence on the urban microclimate. For instance, shade from trees becomes less effective at reducing temperature when they overlap with shade from buildings¹²⁰. Tree or shrub placement can also influence overall ventilation and result in heat or pollutant trapping within urban canyons, with model simulations suggesting up to a 40% increase in pollutant concentration arising from the presence of a dense row of vegetation¹²¹.

Indeed, simulations can be used to understand the impact of plant placement on the microclimate to facilitate planning. The results, however, are often highly contextual and dependent not just on plant attributes but also on the energy exchange characteristics from its surroundings. For instance, while some simulations suggest maximum cooling for trees planted at equal intervals¹²², others indicate that a clustered arrangement provided the largest cooling effect¹⁰⁸ (FIG. 4f). Nevertheless, it is clear that the selection and placement of plants must be considered in a manner that optimizes shading and evapotranspiration, and, thereby, cooling (FIG. 4e).

Greenery on buildings

Modern urban landscapes are characterized by their compact city form, leaving little space for parks and gardens. In this regard, green roofs and vertical greenery (or living walls, green facades) — where vegetation is transplanted onto building surfaces (FIG. 5) — serve as alternatives to traditional landscape, providing environmental benefits to the cityscape, without much demand for ground-level space¹²³. Much like greenery on the ground, greenery on buildings acts to cool by modifying evapotranspiration, shade provision and albedo, but also reduces heat transmission into (and out of) the building envelope by enhancing thermal insulation¹²⁴ (FIG. 2).

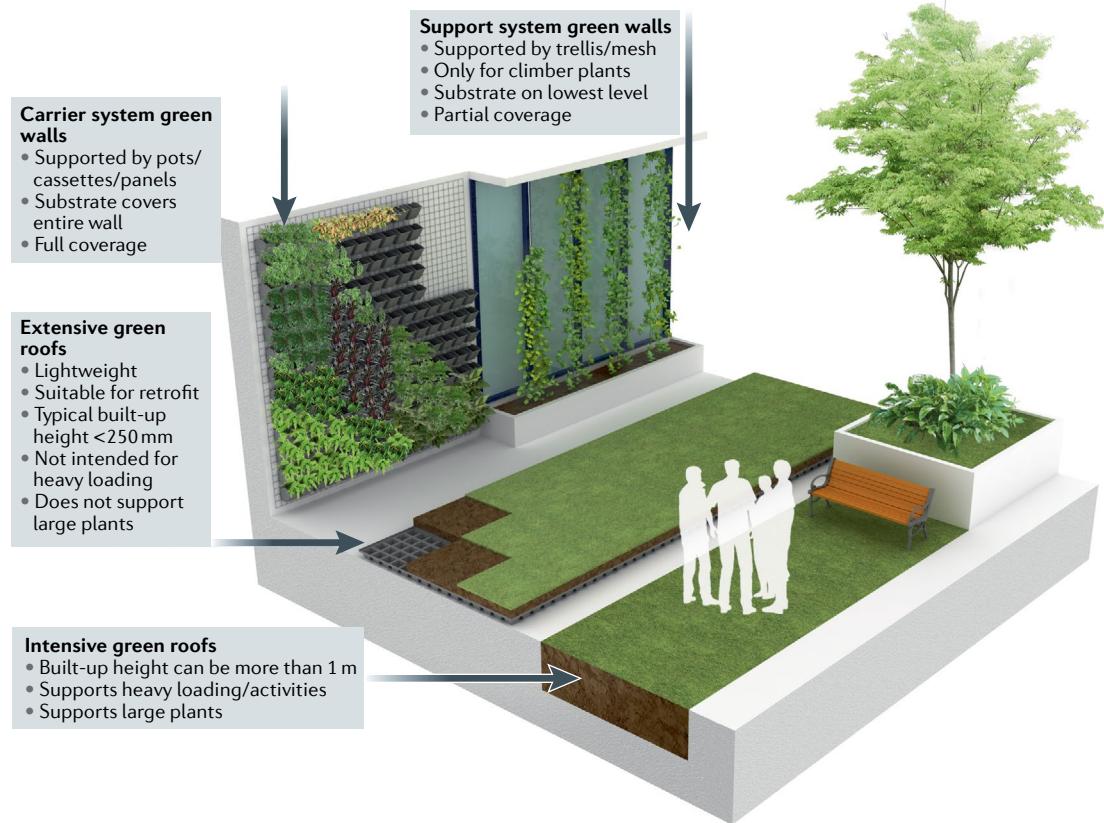


Fig. 5 | Types of greenery on buildings. Different options for vertical and rooftop greenery, including support system green walls, carrier system green walls, extensive green roofs and intensive green roofs.

Such greenery strategies can, therefore, feed back to provide energy savings.

Numerous studies have confirmed the cooling effect of vertical and rooftop greenery, in some cases, reporting surface temperatures up to 20 °C (REFS^{62,125}) and >10 °C (REFS^{39,126,127}) cooler than exposed surfaces, respectively. On average, however, it is estimated that green roofs can reduce peak air temperature by an average of ~3 °C (with a range of 1.5–4.1 °C) and peak surface temperature by an average of ~17 °C (with a range of 11–22.4 °C). Similarly, green walls are able to reduce peak air temperature by an average of ~3 °C (with a range of 2–4 °C) and peak surface temperature by an average of ~16 °C (with a range of 10.7–18.8 °C) (FIG. 3b). As indicated, there are significant differences within these systems that influence cooling potential, as now discussed.

Climate

The magnitude of surface temperature reduction in the presence of vertical and rooftop greenery is highly dependent on prevailing climatic conditions, particularly the season. Typically, both green walls and green roofs are most effective in the summer, when there is higher evapotranspiration and foliage density (thus, shade provision and albedo). For example, green roofs in a Mediterranean climate (Italy) have been shown to reduce peak surface temperature by 20–30 °C in the summer, but only 10–13 °C during winter¹²⁸ (FIG. 6a,b), consistent with other analyses in Shanghai¹²⁹, Estonia¹³⁰ and

Michigan¹³¹. A seasonal analysis for green walls in Italy also shows a similar pattern, wherein reductions in surface temperature peaked at around 6–7 °C in the summer, but only 3.5 °C in winter¹³², with similar trends found in the UK¹³³. Thus, greenery on building surfaces can act to reduce heat gain in summer and abate heat loss in winter (FIG. 6a,b).

As well as the season, the effectiveness of green facades and green roofs is also strongly influenced by the meteorological conditions. In particular, temperature reduction is most effective in sunny weather, becoming less potent during cloudy or rainy periods^{134,135}; in Hong Kong, for instance, green-roof-related maximum surface temperature cooling was 19.8 °C, 7.74 °C and 7.85 °C in sunny, cloudy and rainy weather, respectively¹³⁶. This meteorological sensitivity arises from reduced longwave and shortwave radiation exposure during cloudy periods, curtailing temperature rise and, therefore, cooling potential, and reduced evapotranspiration during periods of high cloud cover or rain, owing to changes in solar irradiance and vapour pressure deficit^{119,137}. Indeed, it is thought that a threshold of approximately 300 W m⁻² must be crossed before evapotranspiration cooling is activated and becomes evident¹³⁸.

System selection and placement

Intensive green roof versus extensive green roof. Green roofs can be intensive or extensive (FIG. 5), respectively encompassing those that are designed for public access

Vapour pressure deficit
The difference between moisture content in in situ air compared with the total moisture the air can hold when it is saturated.

or not, influencing their characteristics; intensive green roofs have deeper soil depths (>250 mm) and can hold large shrubs or small trees^{124,139}, whereas extensive green roofs have shallow soil depths (~150 mm) and planting palettes limited to succulents and shrubs¹⁴⁰. Accordingly, intensive systems exhibit greater heat absorption and reduced temperature fluctuations¹⁴¹. For example, intensive green roofs have been found to reduce surface temperature by 30 °C, but also provide lower air temperature by up to 4.2 °C at 0.3 m height¹²⁵. In comparison, a separate analysis of a green roof consisting of *Sedum* plants recorded temperature reductions of less than 2 °C (REF¹⁴²). With careful plant selection, however, it is still possible to achieve peak air temperature cooling of 4.5 °C (REF¹³⁴).

Support green wall versus carrier green wall. As with green roofs, there are also different types of green walls: carrier system or support system. In the former, the plant substrate is distributed over the entire wall, whereas in the latter, the substrate is limited to the bottom and mesh used to support climber plants (FIG. 5). As such, carrier systems tend to exhibit greater insulation capabilities than support systems, owing to the substructure, air gap, substrate and plant layers^{126,143}. In some cases, differences of close to 11 °C have been documented, with carrier systems and support systems promoting surface cooling of 21.5 °C and 10.7 °C, respectively, in Spain¹⁴⁴. Accordingly, energy savings reached 23% for carrier green walls and 19% for support green walls¹⁴⁴.

System placement. The ability of green walls and green roofs to reduce temperature in the urban environment is contingent upon several morphological factors, such as wall size and shape, as well as conditions of their immediate surroundings. Plant cooling potential from shade and evapotranspiration can be severely undermined when shade is already provided by adjacent buildings or when lack of access to sunlight impedes the evapotranspiration process (transpiration is a light-dependent process)¹⁴⁵. Given that green roofs are typically located high on buildings, the chances of overshadowing from taller structures is often small.

Green walls, by contrast, are often influenced by their placement, specifically, through self-shading and overshadowing from other buildings. In one instance, green-wall-related temperature reductions have dropped from 16 °C to 2 °C as a result of self-shading, illustrating the dependence of time of day on cooling potential¹²⁷. Similarly, the orientation of green walls is also important (FIG. 6c,d): east-facing and west-facing walls typically experience maximum cooling potential at different times of the day, owing to the direction exposure of the rising and setting sun, in one instance, lowering peak surface temperatures by 15 °C and 16.4 °C at around 12:00 and 19:00, respectively¹⁴⁶. The south-facing facade, by contrast, recorded its peak temperature reduction of 16 °C at 15:45 (FIG. 6c).

Plant selection

Plant selection further has a direct impact on the cooling potential of both green walls and green roofs. Selecting plants with large foliage will result in higher shade

provision and less exposure to both longwave and shortwave radiation, lowering surface and air temperature outdoors, as well as reducing heat transmitted into the building. Green roofs with bigger shrubs and deeper soil depths tend to provide better cooling, as there is more shading from the plant canopy^{147,148}.

In addition to foliage density, plant functional attributes such as leaf size and colour, as well as evapotranspiration rate, further contribute to temperature reduction. Plants with high LAI, leaf stomatal conductance (indicating transpiration activity), thin and light leaf colour provide better cooling¹⁴⁹. Specifically, *Stachys byzantina* and *Salvia officinalis*, both tall non-succulent plants with high LAI, can register surface temperature reductions of up to 10 °C. Notably, temperature reduction provided by *Sedum* (succulent, shortest plant and lowest stomatal conductance) was approximately 5 °C.

The holistic approach to plant selection is important, as the cooling potential of plants can vary between species. For instance, plants with high LAI can contribute to higher mean radiant temperature exposure, as larger leaves might get heated up more easily than plants with smaller leaves, as indicated by a 10 °C difference in peak mean radiant temperature between *Phyllanthus cochinchinensis* and *Heliconia 'American Dwarf'*⁶² (FIG. 6e). Therefore, it is more appropriate to select plants based on a variety of traits such as plant height, evapotranspiration rate and albedo, instead of focusing solely on foliage density.

Similar to green roofs, the choice of plants selected will greatly influence the cooling potential of green walls. Variations in plant evapotranspiration and shade provision can account for close to 4 °C in temperature reduction (FIG. 6f). Plants such as *Jasminum officinale* display more cooling due to shade provision, while cooling from *Fuchsia 'Lady Boothby'* can be attributed more to transpiration activity¹⁵⁰.

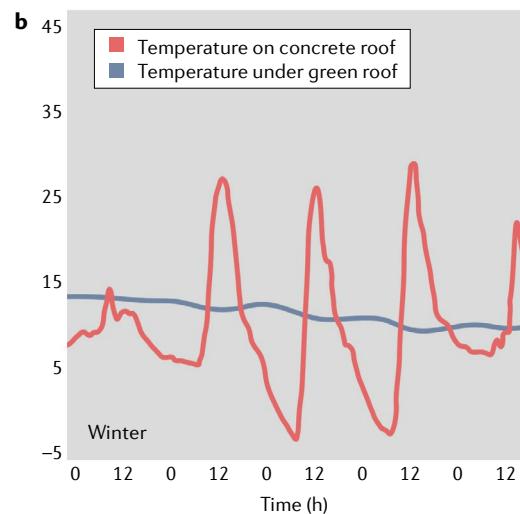
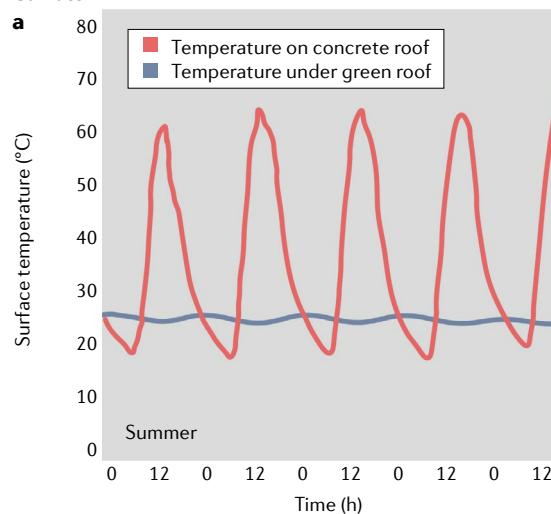
Implications for urban design

It is clear that urban greenery is effective in reducing temperature in the built environment: green parks lower air and surface temperatures by an average of 3 °C and 7 °C, respectively, while green walls and green roofs can reduce peak surface temperature by around 17 °C. Moreover, the variability in cooling potential (FIG. 3) implies that simply adding greenery might not instantly cool temperatures. Instead, a more nuanced approach must be adopted when it comes to the selection and placement of greenery, bearing in mind biotic as well as abiotic considerations (FIGS 4,6). While there are no universally accepted guidelines, several key concepts from this Review can be used to help inform urban design and maximize the mitigation potential of urban greenery.

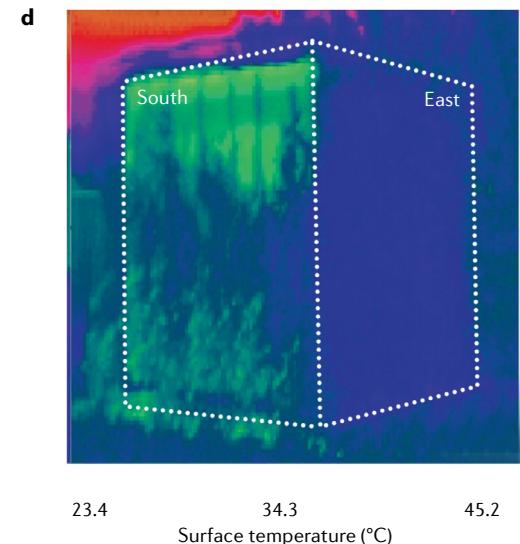
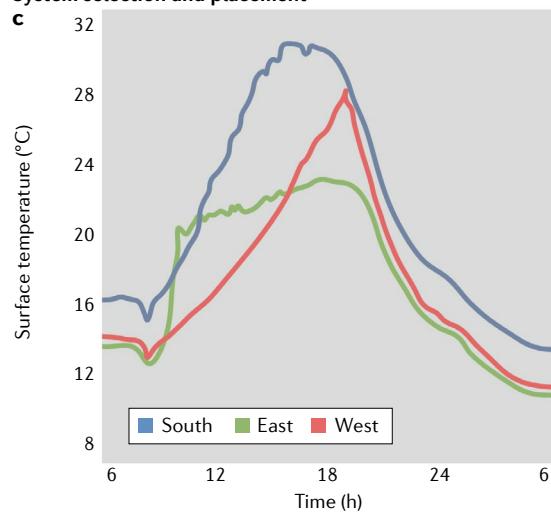
Greenery on ground

There is evidence to suggest that urban parks should be 0.5–1 ha in size to maximize their cooling potential^{97–100}. In addition, they should be regularly shaped to minimize anthropogenic influences and capitalize on the TVoE (FIG. 4c,d). Ideally, these parks should also be evenly interspersed throughout the city with spacing of less than 1 km (REFS^{73,107,151}), all to ensure maximum

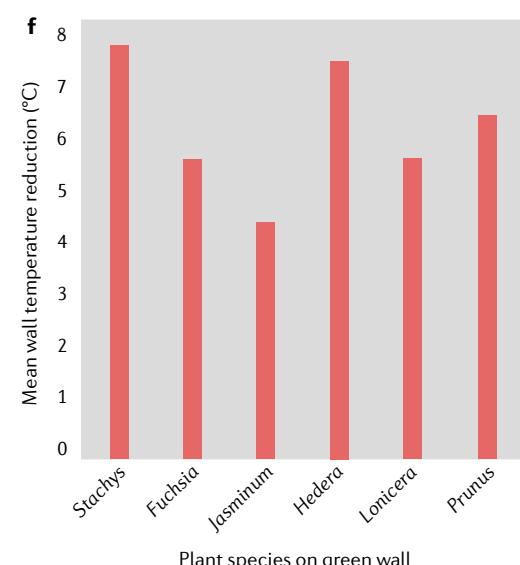
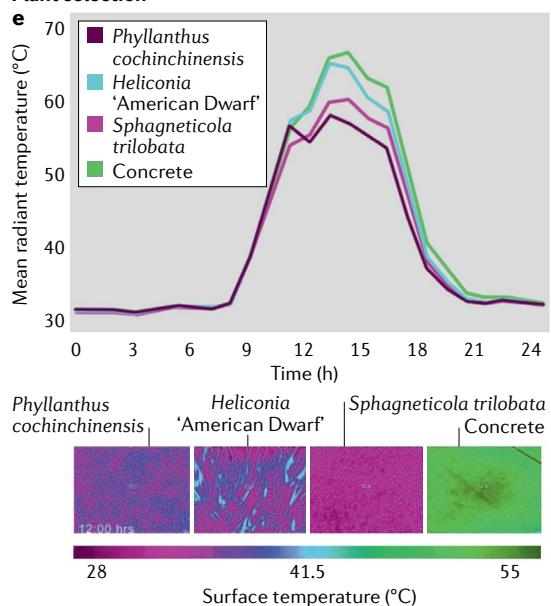
Climate



System selection and placement



Plant selection



◀ Fig. 6 | **Factors influencing the cooling potential of vertical and rooftop greenery.**

Climate: surface temperature of a green roof (blue) and a concrete roof (red) in summer (panel **a**) and winter (panel **b**) in Italy; note the different scales on the y axes. Peak temperatures are reduced to a much larger extent in summer compared with winter, but, in both cases, temperature fluctuations are greatly reduced. System selection and placement: surface temperature profiles associated with green walls on south-facing (blue), east-facing (green) and west-facing (red) surfaces (panel **c**), and an infrared image illustrating the difference between south-facing and east-facing walls (panel **d**). Plant selection: diurnal radiant temperature profiles and corresponding infrared images for three different plant species and concrete (panel **e**), and mean wall surface temperature reduction from six plant species (panel **f**). Panels **a** and **b** are adapted with permission from REF¹²⁸. Panels **c** and **d** are adapted with permission from REF¹⁴⁶. Panel **e** is adapted with permission from REF⁶². Panel **f** is adapted with permission from REF¹⁵⁰.

spillover of cooler air temperatures^{84,152,153}. As much as possible, it is advisable to allocate more trees than lawn space to block more direct solar radiation⁹⁴. For smaller green spaces to provide effective cooling, plants need to be selected and placed strategically, as informed through the use of simulation tools to visualize cooling effect from trees^{108,154} (FIG. 4f).

Pavement or roadside trees are able to reduce surface and air temperature to a lesser degree, but are typically located in close proximity to sources of anthropogenic activity. Their ability to provide cooling is also more susceptible to influences from the urban geometry and climate^{155,156}, providing opportunities to select vegetation with specific traits, such as choosing trees with larger canopies, so that more shade is provided along pedestrian footpaths during the day¹¹⁶. Some examples include *Caesalpinia ferrea* and *Peltophorum pterocarpum* for tropical¹¹⁹, *Handroanthus chrysotrichus* and *Caesalpinia pluviosa* for subtropical¹¹⁵, and *Tilia cordata* for temperature regions¹¹¹.

Several major cities already have existing directives for park provision per capital¹²³, as well as urban master plans to gradually increase green cover. In the current London Plan, greenery is a key component in the themes of 'Creating a healthy city' and 'Increase efficiency and resilience'¹⁵⁷. A target of making more than half of London green by 2050 was set out, with policies implemented to encourage the inclusion of green infrastructure and promote the creation of new publicly accessible green spaces within urban areas, in turn, increasing the city's resilience and adaptation to climate change and the UHI.

Greenery on buildings

Both intensive and extensive green roofs can further help reduce temperature (FIG. 3b). Larger temperature reductions can be expected from intensive green roof systems with deep soil depths and large plants such as *Rhapis excelsa* and *Erythrina fusca*^{125,142}. Similarly, green walls provide thermal benefits by shielding wall surfaces from direct sunlight. Carrier systems are more effective than support systems at reducing heat gain through the building facade. For support systems, substantial coverage can also be achieved over time, but climber plants require sufficient time to grow, as well as proper maintenance for thick and even coverage.

It is clear that site conditions must be determined before commissioning a green wall or green roof project.

It is recommended that solar simulation be conducted first, with adjacent buildings included to account for possible overshadowing effects. Building walls with high solar insolation are suitable for green wall installation. Green walls can still be installed for well-shaded areas, but their impact on temperature reduction will be diminished.

Next, the type of green wall system should be specified. Carrier systems are preferred for solid walls, as they provide better insulation and provide consistent coverage throughout the green wall. Support systems can be used for glazed facades, but design has to take into account factors such as view, access and maintainability. Designers might opt for smaller gaps between trellises and planting multiple climber plants per trellis to provide thicker and more consistent coverage.

Finally, plants need to be selected for maximizing cooling. Plants with big leaves (high LAI) such as *Aristolochia acuminata* are recommended, as they provide more shade and have less risk of being overcrowded compared with plants with smaller leaves, such as *Selaginella* sp.

The latest Singapore Master Plan has set a target of increasing greenery to ensure thermal comfort in light of climate change¹⁵⁸. These targets are supported by legislation such as the Landscape Replacement Policy as well as the BCA Green Mark Scheme, which encourages the adoption of sky-rise greenery (green roofs and green walls) with emphasis on shade provision from plants to reduce the UHI. Greenery density is quantified using the green plot ratio (GnPR) — a function of green space area and corresponding LAI¹⁵⁹ — as a more direct translation of academic knowledge of the benefits of high LAI (leading to higher shade provision and more cooling) into practice.

Elsewhere, the promotion of sky-rise greenery in cities is done either through legislation (such as the 2019 Green Roof Act in New York¹⁶⁰, the Biodiversity Act and Green Roof Statement in France¹⁶¹ and the Tokyo Green Roof Law¹⁶²) or as a criteria in Green Building Rating Tools (GBRTs). Established GBRTs such as LEED (USA) require the installation of extensive or intensive green roofs as part of vegetation provision¹⁶³. In BEAM Plus (Hong Kong), points are allocated for the provision of vegetated building envelope and green roofs to reduce thermal impact¹⁶⁴. Besides ensuring adequate greenery provision via GnPR quantification, the BCA Green Mark Scheme (Singapore) also awards points for advanced greening efforts, such as having green walls on the east-facing and west-facing facades, to reduce direct solar exposure and minimize heat gain¹⁶⁵.

Integration

When considering the combined effects of greenery, the scale of mitigation benefits must be identified¹⁶⁶ (FIG. 7a). At the city scale, improvement of greenery type for existing parks takes priority, while for the installation of new green parks, cooling intensity and cooling effect distance should be quantified using urban spatial modelling tools. At the district scale, street dimension and street canyon characteristics have a major role; street trees and vertical greenery are the most appropriate solutions. At the neighbourhood scale, areas where citizens are exposed

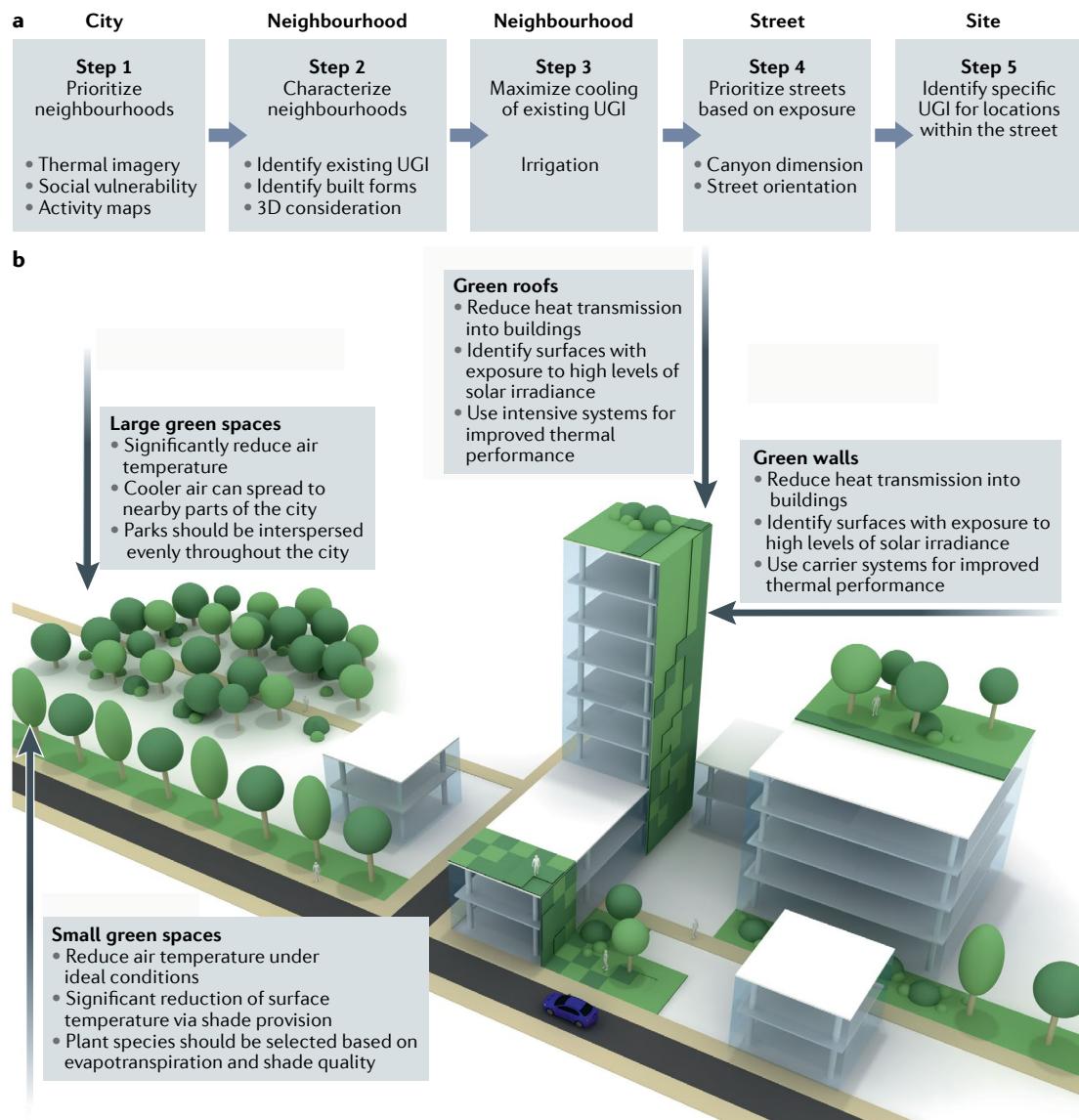


Fig. 7 | Translation of greenery research into design. **a** | Proposed framework for implementing greenery at different scales. **b** | A summary of cooling benefits of urban greenery. UGI, urban green infrastructure. Panel **a** is adapted with permission from REF.¹⁶⁶.

to high mortality and morbidity due to excess heat should be identified. Pocket parks, small green spaces and vertical greenery are the first to be considered, especially if the identified neighbourhoods are densely built and populated.

To better mitigate urban heat, it is advisable to employ a combination of parks, tree and shrub plantings, and vertical and rooftop greenery, situated at areas where building occupants and pedestrians can benefit most from their cooling benefits (FIG. 7b). The different forms of greenery occupy different spaces in the built environment, so they are not mutually exclusive. It is important for the planning of greenery provision to commence at the design stage and not as an afterthought.

Simulation of solar exposure should be conducted for all surfaces (including ground level, as well as the building envelope) for the purpose of identifying areas of high solar insolation, where the addition of greenery

should be prioritized. This addition could be selecting trees with large canopies to shade main pedestrian paths or assigning green walls for facades with high solar exposure. Solar exposure is highly contextual and dependent on the local climate. Therefore, adjacent buildings should also be modelled and appropriate weather files be used for simulating conditions for the entire year. Once vulnerable areas have been identified, designers can start to select the appropriate plant species or greenery systems. After designing with the solar simulation results, architects can conduct iterative simulation studies to determine the impact of their greenery design schemes on thermal comfort and energy savings. This process can be repeated until the desired outcome is achieved. Urban planners can adopt a similar methodology for park design. Green spaces should be evenly interspersed throughout the city, so that cooling from parks can be more widespread. Care should be taken

to have less convoluted site plans and to maximize tree canopy coverage.

Summary and future perspectives

Peak UHI intensities can reach 10 °C (REFS^{5,167}). This Review has shown that greenery, in all its forms, can be used strategically to alleviate heat gain, reduce thermal stress and, thereby, morbidity and mortality. Indeed, through shade provision and evapotranspiration, greenery on the ground and greenery on buildings can reduce air temperatures by ~3 °C (FIG. 3). However, the cooling benefits of urban greenery, be it greenery on the ground or on buildings, is not homogeneous, and is influenced by climate, plant selection and placement, as well as size and shape for green parks (FIGS 4, 6). When rightly translated into design guidelines, scientific understanding of urban greenery can, thus, inform future urban design, which is vital, given anthropogenic climate change and the rising incidence of heatwaves. However, future research is required to maximize the potential of urban vegetation as a mitigation tool, including the following.

Plant functional traits database

While the cooling benefits of different forms of greenery have been widely established, many studies point to the lack of specific data to make informed choices when deciding on the size and shape of green spaces, or when specifying plants for enhancing cooling; that is, there has not been a coordinated effort to come up with a comprehensive database of plant functional trait values at the species level. This database could include growth performance of plants to facilitate green wall coverage¹⁶⁸, to information on drought tolerance of trees for resilient streetscapes in light of changing weather patterns and water availability¹¹⁹.

Owing to the lack of complete information, urban designers are not able to actively select plants that can provide more cooling. More importantly, this absence of information can lead to inaccuracies when simulating the cooling effects of greenery, as input factors such as LAI, evapotranspiration and vegetation coverage^{169,170} are inadvertently generalized¹⁷¹. By having a consolidated database, researchers can build on existing knowledge and minimize testing plant species that have already been tested in previous studies. To ensure robustness of results, standards for setting up experiments can also be recommended, including minimum measurement periods, data logging frequency and specifications for sensor quality. Replicability of tests, which is a critical indication of reliability of methodology but often overlooked in this field of study^{172,173}, can ensue. This can range from growth performance of plants to facilitate green wall coverage¹⁶⁸ to information on drought tolerance of trees for resilient streetscapes in light of changing weather patterns and water availability¹¹⁹.

Physiological equivalent temperature

Air temperature at which, in a typical indoor setting, the heat balance of the human body is maintained with core and skin temperatures equal to those under the conditions being assessed. It provides an indication of thermal comfort, applicable for both indoors and outdoors.

Computational fluid dynamics

(CFD). Quantitative modelling of fluid flow based on the laws of mass, momentum and energy conservation that govern fluid motion.

temperature, air velocity and relative humidity^{174,175}. A single tree, for example, has been found to be able to reduce the physiological equivalent temperature by up to 11 °C, but with marked variability owing to myriad factors, including time of day, surrounding vegetation, urban geometry and prevailing wind flow near the measurement spot¹⁷⁶. Thus, although greenery can significantly improve thermal comfort by reducing the mean radiant temperature through shade provision and evapotranspiration, it is equally possible that trees or tall shrubs impede wind flow, leading to overall thermal discomfort¹⁷⁷. Indeed, other studies have gone further to show that inappropriate tree placement can be highly detrimental to the outdoor environment, hindering anthropogenic heat and pollutant dispersion in the urban environment^{178,179}.

To address these issues, computational fluid dynamics (CFD) simulations can be used to provide some understanding on how placement of vegetation, especially for trees and tall shrubs, can complement prevailing wind conditions^{71,180}. However, given the complexity and computing resources required, understanding of greenery impacts on wind flow using CFD is in its infancy and a resource ready to be fully utilized.

Translation into design guidelines

Much as research is crucial to understanding the cooling benefits of greenery, it is equally important to translate what is known into practical design. These guidelines can range from simple rules such as prioritizing east–west orientations for green wall facings to maximizing heat reduction for the building¹⁴⁶, selecting trees with high canopy density¹⁷², to complex frameworks that take into consideration surrounding built morphology and macroscale and microscale variables¹⁶⁶. Different modes of greenery infrastructure can be recommended based on the corresponding scale of intervention, with consideration of other mitigating factors, such as urban geometry and climate (FIG. 7a). Through this method, specific greenery needs of the site can be addressed adequately and comprehensively in a ‘right tree, right place’ approach^{154,181}. This performance-based (or translational) approach can further look into pedestrian comfort at predefined routes and examine how canopy shapes and tree placement can maximize shade provision, leading to improved thermal comfort.

Besides reducing temperature, the presence of greenery can also improve air quality¹⁸², promote urban biodiversity¹⁸³ and stimulate mental as well as physiological well-being¹⁸⁴. As an ecosystem service, greenery, therefore, offers a plethora of benefits to the urban environment, the information of which must be available to maximize greenery benefits. In addition, these design frameworks provide the opportunity for designers to couple other forms of climate regulation services into their design schema, from blue infrastructure (waterways)³⁷, cool materials (cool roofs and cool pavements)³³, to retroreflective facades¹⁸⁵. Such efforts are key to encouraging the industry to adopt greening practices and improving thermal conditions of the urban environment.

1. Akbari, H. & Kolokotsa, D. Three decades of urban heat islands and mitigation technologies research. *Energy Build.* **133**, 834–842 (2016).
2. Oke, T. R. The energetic basis of the urban heat island. *Q. J. R. Meteorol. Soc.* **108**, 1–24 (1982). **Pioneering work describing and quantifying the urban heat island effect.**
3. Oke, T. R., Johnson, G. T., Steyn, D. G. & Watson, I. D. Simulation of surface urban heat islands under 'ideal' conditions at night part 2: Diagnosis of causation. *Boundary-Layer Meteorol.* **56**, 339–358 (1991).
4. He, X. et al. Observational and modeling study of interactions between urban heat island and heatwave in Beijing. *J. Clean. Prod.* **247**, 119169 (2020).
5. Santamouris, M. Analyzing the heat island magnitude and characteristics in one hundred Asian and Australian cities and regions. *Sci. Total Environ.* **512–513**, 582–598 (2015).
6. Deilami, K., Kamruzzaman, M. & Liu, Y. Urban heat island effect: a systematic review of spatio-temporal factors, data, methods, and mitigation measures. *Int. J. Appl. Earth Obs. Geoinf.* **67**, 30–42 (2018).
7. Rauf, S. et al. How hard they hit? Perception, adaptation and public health implications of heat waves in urban and peri-urban Pakistan. *Environ. Sci. Pollut. Res.* **24**, 10630–10639 (2017).
8. Sakka, A., Santamouris, M., Livada, I., Nicol, F. & Wilson, M. On the thermal performance of low income housing during heat waves. *Energy Build.* **49**, 69–77 (2012).
9. Rizvi, S. H., Alam, K. & Iqbal, M. J. Spatio-temporal variations in urban heat island and its interaction with heat wave. *J. Atmos. Sol. Terr. Phys.* **185**, 50–57 (2019).
10. Founda, D. & Santamouris, M. Synergies between urban heat island and heat waves in Athens (Greece), during an extremely hot summer (2012). *Sci. Rep.* **7**, 10973 (2017).
11. Heaviside, C., Vardoulakis, S. & Cai, X.-M. Attribution of mortality to the urban heat island during heatwaves in the West Midlands, UK. *Environ. Health* **15**, S27 (2016).
12. Robine, J.-M. et al. Death toll exceeded 70,000 in Europe during the summer of 2003. *C. R. Biol.* **331**, 171–178 (2008).
13. Le Tertre, A. et al. Impact of the 2003 heatwave on all-cause mortality in 9 French cities. *Epidemiology* **17**, 75–79 (2006).
14. Tan, J. et al. The urban heat island and its impact on heat waves and human health in Shanghai. *Int. J. Biometeorol.* **54**, 75–84 (2010).
15. Goggins, W. B., Chan, E. Y. Y., Ng, E., Ren, C. & Chen, L. Effect modification of the association between short-term meteorological factors and mortality by urban heat islands in Hong Kong. *PLoS ONE* **7**, e38551 (2012).
16. Dang, T. N., Van, D. Q., Kusaka, H., Seposo, X. T. & Honda, Y. Green space and deaths attributable to the urban heat island effect in Ho Chi Minh City. *Am. J. Public Health* **108**, S137–S143 (2017).
17. Paravantis, J., Santamouris, M., Cartalis, C., Efthymiou, C. & Kontoulis, N. Mortality associated with high ambient temperatures, heatwaves, and the urban heat island in Athens, Greece. *Sustainability* **9**, 606 (2017).
18. Milojevic, A. et al. Impact of London's urban heat island on heat-related mortality. *Epidemiology* **22**, S182–S183 (2011).
19. Santamouris, M. Heat island research in Europe: the state of the art. *Adv. Build. Energy Res.* **1**, 123–150 (2007).
20. Tran, H., Uchihama, D., Ochi, S. & Yasuoka, Y. Assessment with satellite data of the urban heat island effects in Asian mega cities. *Int. J. Appl. Earth Obs. Geoinf.* **8**, 34–48 (2006).
21. Conry, P. et al. Chicago's heat island and climate change: Bridging the scales via dynamical downscaling. *J. Appl. Meteorol. Climatol.* **54**, 1430–1448 (2015).
22. Yang, L. et al. Contrasting impacts of urban forms on the future thermal environment: example of Beijing metropolitan area. *Environ. Res. Lett.* **11**, 034018 (2016).
23. Sachindra, D. A., Ng, A. W. M., Muthukumaran, S. & Perera, B. J. C. Impact of climate change on urban heat island effect and extreme temperatures: a case-study. *Q. J. R. Meteorol. Soc.* **142**, 172–186 (2016).
24. Lemonsu, A., Kounkou-Arnaud, R., Desplat, J., Salagnac, J.-L. & Masson, V. Evolution of the Parisian urban climate under a global changing climate. *Clim. Change* **116**, 679–692 (2013).
25. Lauwaet, D. et al. Assessing the current and future urban heat island of Brussels. *Urban Clim.* **15**, 1–15 (2016).
26. Chapman, S., Watson, J. E. M., Salazar, A., Thatcher, M. & McAlpine, C. A. The impact of urbanization and climate change on urban temperatures: a systematic review. *Landscape Ecol.* **32**, 1921–1935 (2017).
27. Li, D. et al. Contrasting responses of urban and rural surface energy budgets to heat waves explain synergies between urban heat islands and heat waves. *Environ. Res. Lett.* **10**, 054009 (2015).
28. Argüeso, D., Evans, J. P., Pitman, A. J. & Di Luca, A. Effects of city expansion on heat stress under climate change conditions. *PLoS ONE* **10**, e0117066 (2015).
29. Li, D. & Bou-Zeid, E. Synergistic interactions between urban heat islands and heat waves: the impact in cities is larger than the sum of its parts. *J. Appl. Meteorol. Climatol.* **52**, 2051–2064 (2013).
30. Takane, Y., Ohashi, Y., Grimmond, C. S. B., Hara, M. & Kikugawa, Y. Asian megacity heat stress under future climate scenarios: impact of air-conditioning feedback. *Environ. Res. Commun.* **2**, 015004 (2020).
31. Lin, C.-Y., Chien, Y.-Y., Su, C.-J., Kueh, M.-T. & Lung, S.-C. Climate variability of heat wave and projection of warming scenario in Taiwan. *Clim. Change* **145**, 305–320 (2017).
32. Taha, H. Urban climates and heat islands: albedo, evapotranspiration, and anthropogenic heat. *Energy Build.* **25**, 99–103 (1997).
33. Doulos, L., Santamouris, M. & Livada, I. Passive cooling of outdoor urban spaces. The role of materials. *Sol. Energy* **77**, 231–249 (2004).
34. Compagnon, R. Solar and daylight availability in the urban fabric. *Energy Build.* **36**, 321–328 (2004).
35. Ratti, C., Di Sabatino, S. & Britter, R. Urban texture analysis with image processing techniques: winds and dispersion. *Theor. Appl. Climatol.* **84**, 77–90 (2006).
36. Sailor, D. J. A review of methods for estimating anthropogenic heat and moisture emissions in the urban environment. *Int. J. Climatol.* **31**, 189–199 (2011).
37. Lin, Y. et al. Water as an urban heat sink: Blue infrastructure alleviates urban heat island effect in mega-city agglomeration. *J. Clean. Prod.* **262**, 121411 (2020).
38. Bowler, D. E., Buying-Ali, L., Knight, T. M. & Pullin, A. S. Urban greening to cool towns and cities: A systematic review of the empirical evidence. *Landscape Urban Plan.* **97**, 147–155 (2010). **Consolidates multiple studies to quantify park cooling effects.**
39. Besir, A. B. & Cuce, E. Green roofs and facades: A comprehensive review. *Renew. Sustain. Energy Rev.* **82**, 915–939 (2018).
40. Ismail, A., Abdul Samad, M. H., Rahman, A. M. A. & Yeoh, F. S. Cooling Potentials and CO₂ uptake of Ipomoea Pes-caprae installed on the flat roof of a single storey residential building in Malaysia. *Procedia Soc. Behav. Sci.* **35**, 361–368 (2012).
41. Nikolić, M. & Stević, S. Family Asteraceae as a sustainable planning tool in phytoremediation and its relevance in urban areas. *Urban For. Urban Green.* **14**, 782–789 (2015).
42. Lin, M.-Y. et al. The effects of vegetation barriers on near-road ultrafine particle number and carbon monoxide concentrations. *Sci. Total Environ.* **553**, 372–379 (2016).
43. Cook-Patton, S. C., McArt, S. H., Parachnowitsch, A. L., Thaler, J. S. & Agrawal, A. A. A direct comparison of the consequences of plant genotypic and species diversity on communities and ecosystem function. *Ecology* **92**, 915–923 (2011).
44. Menz, M. H. M. et al. Reconnecting plants and pollinators: challenges in the restoration of pollination mutualisms. *Trends Plant Sci.* **16**, 4–12 (2011).
45. Takebayashi, H. & Moriyama, M. Surface heat budget on green roof and high reflection roof for mitigation of urban heat island. *Build. Environ.* **42**, 2971–2979 (2007).
46. Hoyano, A. Climatological uses of plants for solar control and the effects on the thermal environment of a building. *Energy Build.* **11**, 181–199 (1988).
47. Taha, H. in *Analysis of Energy Efficiency of Air Quality in the South Coast Air Basin-Phase II, Report No. LBL-35728* (ed. Taha, H. et al.) 43–59 (Lawrence Berkeley National Laboratory, 1994).
48. Tan, P. Y. et al. A method to partition the relative effects of evaporative cooling and shading on air temperature within vegetation canopy. *J. Urban Ecol.* **4**, juy012 (2018).
49. Hoelscher, M.-T., Nehls, T., Jänicke, B. & Wessolek, G. Quantifying cooling effects of facade greening: shading, transpiration and insulation. *Energy Build.* **114**, 283–290 (2016).
50. Papadakis, G., Tsamis, P. & Kyritsis, S. An experimental investigation of the effect of shading with plants for solar control of buildings. *Energy Build.* **33**, 831–836 (2001).
51. Simpson, J. R. Improved estimates of tree-shade effects on residential energy use. *Energy Build.* **34**, 1067–1076 (2002).
52. Heisler, G. M. Energy savings with trees. *J. Arboricul.* **12**, 113–125 (1986).
53. McPherson, E. G., Herrington, L. P. & Heisler, G. M. Impacts of vegetation on residential heating and cooling. *Energy Build.* **12**, 41–51 (1988).
54. McPherson, E. G., Simpson, J. R. & Livingston, M. Effects of three landscape treatments on residential energy and water use in Tucson, Arizona. *Energy Build.* **13**, 127–138 (1989).
55. Parker, J. H. Landscaping to reduce the energy used in cooling buildings. *J. Forestry* **81**, 82–105 (1983).
56. Oke, T. R. *Boundary Layer Climates* (Routledge, 2002).
57. Seyam, S. The impact of greenery systems on building energy: systematic review. *J. Build. Eng.* **26**, 100887 (2019).
58. He, Y., Yu, H., Ozaki, A., Dong, N. & Zheng, S. Influence of plant and soil layer on energy balance and thermal performance of green roof system. *Energy* **141**, 1285–1299 (2017).
59. Cleugh, H. & Grimmond, S. in *The Future of the World's Climate* 2nd edn (eds Henderson-Sellers, A. & McGuffie, K. E.) 47–76 (Elsevier, 2011).
60. Dabbert, W. F. & Davis, P. A. Determination of energetic characteristics of urban-rural surfaces in the greater St. Louis area. *Boundary-Layer Meteorol.* **14**, 105–121 (1978).
61. Steyn, D. & Oke, T. Effects of a small scrub fire on the surface radiation budget. *Weather* **35**, 212–215 (1980).
62. Tan, C. L., Wong, N. H., Tan, P. Y., Jusuf, S. K. & Chiam, Z. Q. Impact of plant evapotranspiration rate and shrub albedo on temperature reduction in the tropical outdoor environment. *Build. Environ.* **94**, 206–217 (2015). **Quantifies plant traits for green roof shrubs and their impact on cooling.**
63. Dobos, E. in *Encyclopedia of Natural Resources-Land Vol. I* (ed. Wang, Y.) 7–9 (CRC Press, 2014).
64. Skoulika, F., Santamouris, M., Kolokotsa, D. & Boenii, N. On the thermal characteristics and the mitigation potential of a medium size urban park in Athens, Greece. *Landscape Urban Plan.* **123**, 73–86 (2014).
65. Cheung, P. K. & Jim, C. Y. Differential cooling effects of landscape parameters in humid-subtropical urban parks. *Landscape Urban Plan.* **192**, 103651 (2019).
66. Wang, Y., Ni, Z., Peng, Y. & Xia, B. Local variation of outdoor thermal comfort in different urban green spaces in Guangzhou, a subtropical city in South China. *Urban For. Urban Green.* **32**, 99–112 (2018).
67. Oliveira, S., Andrade, H. & Vaz, T. The cooling effect of green spaces as a contribution to the mitigation of urban heat: a case study in Lisbon. *Build. Environ.* **46**, 2186–2194 (2011).
68. Yu, C. & Hien, W. N. Thermal benefits of city parks. *Energy Build.* **38**, 105–120 (2006).
69. Zouli, I., Santamouris, M. & Dimoudi, A. Monitoring the effect of urban green areas on the heat island in Athens. *Environ. Monit. Assess.* **156**, 275 (2008).
70. Tsoka, S., Tsikaloudaki, A. & Theodosiou, T. Analyzing the ENVI-met microclimate model's performance and assessing cool materials and urban vegetation applications—A review. *Sustain. Cities Soc.* **43**, 55–76 (2018).
71. Yang, A.-S., Juan, Y.-H., Wen, C.-Y. & Chang, C.-J. Numerical simulation of cooling effect of vegetation enhancement in a subtropical urban park. *Appl. Energy* **192**, 178–200 (2017).
72. Gromke, C. et al. CFD analysis of transpiration cooling by vegetation: Case study for specific meteorological conditions during a heat wave in Arnhem, Netherlands. *Build. Environ.* **83**, 11–26 (2015).
73. Lin, W., Yu, T., Chang, X., Wu, W. & Zhang, Y. Calculating cooling extents of green parks using remote sensing: method and test. *Landscape Urban Plan.* **134**, 66–75 (2015).
74. Feyisa, G. L., Dons, K. & Meilby, H. Efficiency of parks in mitigating urban heat island effect: an example from Addis Ababa. *Landscape Urban Plan.* **123**, 87–95 (2014).

75. Yu, Z., Guo, X., Jørgensen, G. & Vejre, H. How can urban green spaces be planned for climate adaptation in subtropical cities? *Ecol. Indic.* **82**, 152–162 (2017).
76. Cao, X., Onishi, A., Chen, J. & Imura, H. Quantifying the cool island intensity of urban parks using ASTER and IKONOS data. *Landsc. Urban Plan.* **96**, 224–231 (2010).
77. Saaroni, H., Amorim, J. H., Hiemstra, J. A. & Pearlmutter, D. Urban Green Infrastructure as a tool for urban heat mitigation: Survey of research methodologies and findings across different climatic regions. *Urban Clim.* **24**, 94–110 (2018).
78. Ren, Z. et al. Estimation of the relationship between urban park characteristics and park cool island intensity by remote sensing data and field measurement. *Forests* **4**, 868–886 (2013).
79. Upmanis, H., Eliasson, I. & Lindqvist, S. The influence of green areas on nocturnal temperatures in a high latitude city (Göteborg, Sweden). *Int. J. Climatol.* **18**, 681–700 (1998).
80. Sugawara, H. et al. Thermal influence of a large green space on a hot urban environment. *J. Environ. Qual.* **45**, 125–133 (2016).
81. Nichol, J. Remote sensing of urban heat islands by day and night. *Photogramm. Eng. Remote Sens.* **71**, 613–621 (2005).
82. Hamada, S. & Ohta, T. Seasonal variations in the cooling effect of urban green areas on surrounding urban areas. *Urban For. Urban Green.* **9**, 15–24 (2010).
83. Wong, N. H. & Yu, C. Study of green areas and urban heat island in a tropical city. *Habitat Int.* **29**, 547–558 (2005).
84. Ng, E., Chen, L., Wang, Y. & Yuan, C. A study on the cooling effects of greening in a high-density city: an experience from Hong Kong. *Build. Environ.* **47**, 256–271 (2012).
85. Konarska, J., Holmer, B., Lindberg, F. & Thorsson, S. Influence of vegetation and building geometry on the spatial variations of air temperature and cooling rates in a high-latitude city. *Int. J. Climatol.* **36**, 2379–2395 (2016).
86. Aflaki, A. et al. Urban heat island mitigation strategies: a state-of-the-art review on Kuala Lumpur, Singapore and Hong Kong. *Cities* **62**, 131–145 (2017).
87. Shashua-Bar, L., Pearlmutter, D. & Erell, E. The cooling efficiency of urban landscape strategies in a hot dry climate. *Landsc. Urban Plan.* **92**, 179–186 (2009).
88. Zhao, C., Fu, G., Liu, X. & Fu, F. Urban planning indicators, morphology and climate indicators: A case study for a north-south transect of Beijing, China. *Build. Environ.* **46**, 1174–1183 (2011).
89. Honjo, T. & Takakura, T. Simulation of thermal effects of urban green areas on their surrounding areas. *Energy Build.* **15**, 443–446 (1990).
90. Takebayashi, H. Influence of urban green area on air temperature of surrounding built-up area. *Climate* **5**, 60 (2017).
91. Yan, H., Wu, F. & Dong, L. Influence of a large urban park on the local urban thermal environment. *Sci. Total Environ.* **622–623**, 882–891 (2018).
92. Xiao, X. D., Dong, L., Yan, H., Yang, N. & Xiong, Y. The influence of the spatial characteristics of urban green space on the urban heat island effect in Suzhou Industrial Park. *Sustain. Cities Soc.* **40**, 428–439 (2018).
93. Yu, Z., Guo, X., Zeng, Y., Koga, M. & Vejre, H. Variations in land surface temperature and cooling efficiency of green space in rapid urbanization: The case of Fuzhou city, China. *Urban For. Urban Green.* **29**, 113–121 (2018).
94. Chang, C.-R., Li, M.-H. & Chang, S.-D. A preliminary study on the local cool-island intensity of Taipei city parks. *Landsc. Urban Plan.* **80**, 386–395 (2007).
95. Jaganmohan, M., Knapp, S., Buchmann, C. M. & Schwarz, N. The bigger, the better? The influence of urban green space design on cooling effects for residential areas. *J. Environ. Qual.* **45**, 134–145 (2016).
96. Lu, J., Li, C.-d., Yang, Y.-c., Zhang, X.-h. & Jin, M. Quantitative evaluation of urban park cool island factors in mountain city. *J. Cent. South Univ.* **19**, 1657–1662 (2012).
97. Yu, Z. et al. Critical review on the cooling effect of urban blue-green space: a threshold-size perspective. *Urban For. Urban Green.* **49**, 126630 (2020).
98. Yang, G., Yu, Z., Jørgensen, G. & Vejre, H. How can urban blue-green space be planned for climate adaption in high-latitude cities? A seasonal perspective. *Sustain. Cities Soc.* **53**, 101932 (2020).
99. Yu, Z., Xu, S., Zhang, Y., Jørgensen, G. & Vejre, H. Strong contributions of local background climate to the cooling effect of urban green vegetation. *Sci. Rep.* **8**, 6798 (2018).
100. Fan, H. et al. How to cool hot-humid (Asian) cities with urban trees? An optimal landscape size perspective. *Agric. For. Meteorol.* **265**, 338–348 (2019).
101. Motazedian, A., Coutts, A. M. & Tapper, N. J. The microclimatic interaction of a small urban park in central Melbourne with its surrounding urban environment during heat events. *Urban For. Urban Green.* **52**, 126688 (2020).
102. Du, H. et al. Quantifying the cool island effects of urban green spaces using remote sensing data. *Urban For. Urban Green.* **27**, 24–31 (2017).
103. Park, J., Kim, J.-H., Lee, D. K., Park, C. Y. & Jeong, S. G. The influence of small green space type and structure at the street level on urban heat island mitigation. *Urban For. Urban Green.* **21**, 203–212 (2017).
104. Chen, A., Yao, X. A., Sun, R. & Chen, L. Effect of urban green patterns on surface urban cool islands and its seasonal variations. *Urban For. Urban Green.* **13**, 646–654 (2014).
105. Kato, T., Yamada, T. & Hino, M. Spatial structure of air temperature and humidity in urban park forest and its surrounding. *J. Inst. Sci. Eng. Chuo Univ.* **12**, 63–71 (2006).
106. Moriyama, M., Kono, H., Yoshida, A., Miyazaki, H. & Takebayashi, H. Data analysis on 'cool spot' effect of green canopy in urban areas. *J. Architect. Plan. Environ.* **541**, 49–56 (2001).
107. Vaz Monteiro, M., Doick, K. J., Handley, P. & Peace, A. The impact of greenspace size on the extent of local nocturnal air temperature cooling in London. *Urban For. Urban Green.* **16**, 160–169 (2016).
- Examines the cooling effect beyond park boundaries.**
108. Sodoudi, S., Zhang, H., Chi, X., Müller, F. & Li, H. The influence of spatial configuration of green areas on microclimate and thermal comfort. *Urban For. Urban Green.* **34**, 85–96 (2018).
109. Lin, T.-P., Tsai, K.-T., Hwang, R.-L. & Matzarakis, A. Quantification of the effect of thermal indices and sky view factor on park attendance. *Landsc. Urban Plan.* **107**, 137–146 (2012).
110. Lee, H., Mayer, H. & Chen, L. Contribution of trees and grasslands to the mitigation of human heat stress in a residential district of Freiburg, Southwest Germany. *Landsc. Urban Plan.* **148**, 37–50 (2016).
111. Rahman, M. A., Moser, A., Gold, A., Rötzer, T. & Pauleit, S. Vertical air temperature gradients under the shade of two contrasting urban tree species during different types of summer days. *Sci. Total Environ.* **633**, 100–111 (2018).
112. Kotzen, B. An investigation of shade under six different tree species of the Negev desert towards their potential use for enhancing micro-climatic conditions in landscape architectural development. *J. Arid Environ.* **55**, 231–274 (2003).
113. Lin, B.-S. & Lin, Y.-J. Cooling effect of shade trees with different characteristics in a subtropical urban park. *HortScience* **45**, 83–86 (2010).
114. Berry, R., Livesley, S. J. & Aye, L. Tree canopy shade impacts on solar irradiance received by building walls and their surface temperature. *Build. Environ.* **69**, 91–100 (2013).
115. de Abreu-Harbich, L. V., Labaki, L. C. & Matzarakis, A. Effect of tree planting design and tree species on human thermal comfort in the tropics. *Landsc. Urban Plan.* **138**, 99–109 (2015).
116. Arsmo, D., Stringer, P. & Ennos, A. R. The effect of tree shade and grass on surface and globe temperatures in an urban area. *Urban For. Urban Green.* **11**, 245–255 (2012).
117. Konarska, J., Lindberg, F., Larsson, A., Thorsson, S. & Holmer, B. Transmissivity of solar radiation through crowns of single urban trees — application for outdoor thermal comfort modelling. *Theor. Appl. Climatol.* **117**, 363–376 (2014).
118. Moss, J. L., Doick, K. J., Smith, S. & Shahrestani, M. Influence of evaporative cooling by urban forests on cooling demand in cities. *Urban For. Urban Green.* **37**, 65–73 (2019).
119. Tan, P. Y. et al. Transpiration and cooling potential of tropical urban trees from different native habitats. *Sci. Total Environ.* **705**, 135764 (2020).
120. Thom, J. K., Coutts, A. M., Broadbent, A. M. & Tapper, N. J. The influence of increasing tree cover on mean radiant temperature across a mixed development suburb in Adelaide, Australia. *Urban For. Urban Green.* **20**, 233–242 (2016).
121. Balczó, M., Gromke, C. & Ruck, B. Numerical modeling of flow and pollutant dispersion in street canyons with tree planting. *Meteorol. Z.* **18**, 197–206 (2009).
122. Zhao, Q., Sailor, D. J. & Wentz, E. A. Impact of tree locations and arrangements on outdoor microclimates and human thermal comfort in an urban residential environment. *Urban For. Urban Green.* **32**, 81–91 (2018).
123. Tan, P. Y., Wang, J. & Sia, A. Perspectives on five decades of the urban greening of Singapore. *Cities* **32**, 24–32 (2013).
- Outlines urban greening policies for high-density urban environments.**
124. Vijayaraghavan, K. Green roofs: A critical review on the role of components, benefits, limitations and trends. *Renew. Sustain. Energy Rev.* **57**, 740–752 (2016).
125. Wong, N. H., Chen, Y., Ong, C. L. & Sia, A. Investigation of thermal benefits of rooftop garden in the tropical environment. *Build. Environ.* **38**, 261–270 (2003).
126. Wong, N. H. et al. Thermal evaluation of vertical greenery systems for building walls. *Build. Environ.* **45**, 663–672 (2010).
- One of the first studies to conduct measurements of green walls custom-made for experimentation.**
127. Tan, C. L., Wong, N. H. & Jusuf, S. K. Effects of vertical greenery on mean radiant temperature in the tropical urban environment. *Landsc. Urban Plan.* **127**, 52–64 (2014).
128. Bevilacqua, P., Mazzeo, D., Bruno, R. & Arcuri, N. Experimental investigation of the thermal performances of an extensive green roof in the Mediterranean area. *Energy Build.* **122**, 63–79 (2016).
129. He, Y., Yu, H., Ozaki, A. & Dong, N. Thermal and energy performance of green roof and cool roof: A comparison study in Shanghai area. *J. Clean. Prod.* **267**, 122205 (2020).
130. Teemuisk, A. & Mander, Ü. Greenroof potential to reduce temperature fluctuations of a roof membrane: a case study from Estonia. *Build. Environ.* **44**, 643–650 (2009).
131. Getter, K. L., Rowe, D. B., Andresen, J. A. & Wichman, I. S. Seasonal heat flux properties of an extensive green roof in a Midwestern US climate. *Energy Build.* **43**, 3548–3557 (2011).
132. Vox, G., Blanco, I. & Schettini, E. Green façades to control wall surface temperature in buildings. *Build. Environ.* **129**, 154–166 (2018).
133. Sternberg, T., Viles, H. & Cathersides, A. Evaluating the role of ivy (*Hedera helix*) in moderating wall surface microclimates and contributing to the bioprotection of historic buildings. *Build. Environ.* **46**, 293–297 (2011).
134. Jim, C. Y. & Peng, L. L. H. Weather effect on thermal and energy performance of an extensive tropical green roof. *Urban For. Urban Green.* **11**, 73–85 (2012).
135. Lee, L. S. H. & Jim, C. Y. Thermal-irradiance behaviours of subtropical intensive green roof in winter and landscape-soil design implications. *Energy Build.* **209**, 109692 (2020).
136. Lee, L. S. H. & Jim, C. Y. Thermal-cooling performance of subtropical green roof with deep substrate and woodland vegetation. *Ecol. Eng.* **119**, 8–18 (2018).
137. Cascone, S., Coma, J., Gagliano, A. & Pérez, G. The evapotranspiration process in green roofs: a review. *Build. Environ.* **147**, 337–355 (2019).
138. Jim, C. Y. Thermal performance of climber greenwalls: effects of solar irradiance and orientation. *Appl. Energy* **154**, 631–643 (2015).
139. Kotsiris, G., Nektarios, P. A., Ntoulas, N. & Kargas, G. An adaptive approach to intensive green roofs in the Mediterranean climatic region. *Urban For. Urban Green.* **12**, 380–392 (2013).
140. Skinner, C. J. Urban density, meteorology and rooftops. *Urban Policy Res.* **24**, 355–367 (2006).
141. Jim, C. Y. & Tsang, S. W. Biophysical properties and thermal performance of an intensive green roof. *Build. Environ.* **46**, 1263–1274 (2011).
142. Yin, H., Kong, F., Drnova, I., Middel, A. & James, P. Investigation of extensive green roof outdoor spatio-temporal thermal performance during summer in a subtropical monsoon climate. *Sci. Total Environ.* **696**, 133976 (2019).
143. Wong, N. H., Tan, A. Y. K., Tan, P. Y. & Wong, N. C. Energy simulation of vertical greenery systems. *Energy Build.* **41**, 1401–1408 (2009).
144. Coma, J. et al. Vertical greenery systems for energy savings in buildings: a comparative study between green walls and green facades. *Build. Environ.* **111**, 228–237 (2017).

145. Hohmann-Marriott, M. F. & Blankenship, R. E. Evolution of photosynthesis. *Annu. Rev. Plant Biol.* **62**, 515–548 (2011).
146. Pérez, G., Coma, J., Sol, S. & Cabeza, L. F. Green facade for energy savings in buildings: the influence of leaf area index and facade orientation on the shadow effect. *Appl. Energy* **187**, 424–437 (2017).
147. Saadatian, O. et al. A review of energy aspects of green roofs. *Renew. Sustain. Energy Rev.* **23**, 155–168 (2013).
148. Sailor, D. J., Elley, T. B. & Gibson, M. Exploring the building energy impacts of green roof design decisions – a modeling study of buildings in four distinct climates. *J. Build. Phys.* **35**, 372–391 (2012).
149. Vaz Monteiro, M. et al. Functional green roofs: importance of plant choice in maximising summertime environmental cooling and substrate insulation potential. *Energy Build.* **141**, 56–68 (2017).
150. Cameron, R. W. F., Taylor, J. E. & Emmett, M. R. What's 'cool' in the world of green façades? How plant choice influences the cooling properties of green walls. *Build. Environ.* **73**, 198–207 (2014).
- Quantifies plant traits for green walls and their corresponding cooling effect.**
151. Qiu, K. & Jia, B. The roles of landscape both inside the park and the surroundings in park cooling effect. *Sustain. Cities Soc.* **52**, 101864 (2020).
152. Jamei, E., Rajagopalan, P., Seyedmahmoudian, M. & Jamei, Y. Review on the impact of urban geometry and pedestrian level greening on outdoor thermal comfort. *Renew. Sustain. Energy Rev.* **54**, 1002–1017 (2016).
153. Giridharan, R., Lau, S. S. Y., Ganesan, S. & Givoni, B. Lowering the outdoor temperature in high-rise high-density residential developments of coastal Hong Kong: the vegetation influence. *Build. Environ.* **43**, 1583–1595 (2008).
154. Langenheim, N., White, M., Tapper, N., Livesley, S. J. & Ramirez-Lovering, D. Right tree, right place, right time: a visual-functional design approach to select and place trees for optimal shade benefit to commuting pedestrians. *Sustain. Cities Soc.* **52**, 101816 (2020).
155. Nordh, H. & Ostby, K. Pocket parks for people – a study of park design and use. *Urban For. Urban Green.* **12**, 12–17 (2013).
156. Lin, P., Lau, S. S. Y., Qin, H. & Gou, Z. Effects of urban planning indicators on urban heat island: a case study of pocket parks in high-rise high-density environment. *Landsc. Urban Plan.* **168**, 48–60 (2017).
157. Mayor of London. The London plan. Spatial development strategy for Greater London. *Greater London Authority* <https://www.london.gov.uk/what-we-do/planning/london-plan/new-london-plan/intend-publish-london-plan-2019> (2019).
158. Urban Redevelopment Authority. Singapore master plan. *URA* <https://www.ura.gov.sg/Corporate/Planning/Master-Plan> (2019).
159. Ong, B. L. Green plot ratio: an ecological measure for architecture and urban planning. *Landsc. Urban Plan.* **63**, 197–211 (2003).
160. Espinal, R. L. Jr et al. A local law to amend the administrative code of the city of New York and the New York city building code, in relation to requiring that the roofs of certain buildings be covered in green roofs or solar photovoltaic electricity generating systems. *The New York City Council* <https://legistar.council.nyc.gov/LegislationDetail.aspx?ID=3557657&GUID=B4C3A822-2FBB-45FD-8A74-C59DD95246C1&Options=ID%7cText%7c&Search=1032> (2019).
161. United Nations Framework Convention on Climate Change. France mandates green roofs. *UNFCCC* <https://unfccc.int/news/france-mandates-green-roofs> (2015).
162. Legislative Council Secretariat. Environmental issues in Tokyo (LegCo, 2006).
163. US Green Building Council. LEED public policies. *USGBC* <https://s3.amazonaws.com/legacy.usgbc.org/usgbc/docs/Archive/General/Docs691.pdf> (2010).
164. Hong Kong Green Building Council (HKGBC). BEAM Plus new buildings version 2.0. *HKGBC* https://www.hkgbc.org.hk/eng/beam-plus/file/BEAMPlus_New_Buildings_v2_0.pdf (2019).
165. Building Construction Authority. Green mark for non-residential buildings (GM NRB: 2015) (BCA, 2016).
166. Norton, B. A. et al. Planning for cooler cities: a framework to prioritise green infrastructure to mitigate high temperatures in urban landscapes. *Landsc. Urban Plan.* **134**, 127–138 (2015).
167. Santamouris, M., Cartalis, C., Synnefa, A. & Kolokotsa, D. On the impact of urban heat island and global warming on the power demand and electricity consumption of buildings — a review. *Energy Build.* **98**, 119–124 (2015).
168. Jim, C. Y. Assessing growth performance and deficiency of climber species on tropical greenwalls. *Landsc. Urban Plan.* **137**, 107–121 (2015).
169. Chen, X. et al. Canopy transpiration and its cooling effect of three urban tree species in a subtropical city: Guangzhou, China. *Urban For. Urban Green.* **43**, 126368 (2019).
170. von Arx, G., Graf Pannatier, E., Thimonier, A. & Rebetez, M. Microclimate in forests with varying leaf area index and soil moisture: potential implications for seedling establishment in a changing climate. *J. Ecol.* **101**, 1201–1213 (2013).
171. Peri, G., Rizzo, G., Scaccianoce, G., La Gennusa, M. & Jones, P. Vegetation and soil – related parameters for computing solar radiation exchanges within green roofs: are the available values adequate for an easy modeling of their thermal behavior? *Energy Build.* **129**, 535–548 (2016).
172. Rahman, M. A. et al. Traits of trees for cooling urban heat islands: a meta-analysis. *Build. Environ.* **170**, 106606 (2020).
- A meta-analysis examining tree functional traits for improved cooling potential.**
173. Santamouris, M. et al. Progress in urban greenery mitigation science—assessment methodologies advanced technologies and impact on cities. *J. Civ. Eng. Manag.* **24**, 638–671 (2018).
- A comprehensive review of urban greenery research trends.**
174. American Society of Heating, Refrigerating and Air-Conditioning Engineers. Standard 55 – Thermal environmental conditions for human occupancy (ASHRAE, 2010).
175. Bröde, P. et al. Deriving the operational procedure for the universal thermal climate index (UTCI). *Int. J. Biometeorol.* **56**, 481–494 (2012).
176. Rahman, M. A. et al. Tree cooling effects and human thermal comfort under contrasting species and sites. *Agric. For. Meteorol.* **287**, 107947 (2020).
177. Hami, A., Abdi, B., Zarehaghi, D. & Maulan, S. B. Assessing the thermal comfort effects of green spaces: A systematic review of methods, parameters, and plants' attributes. *Sustain. Cities Soc.* **49**, 101634 (2019).
178. Moradpour, M., Afshin, H. & Farhanieh, B. A numerical investigation of reactive air pollutant dispersion in urban street canyons with tree planting. *Atmos. Pollut. Res.* **8**, 253–266 (2017).
179. Hsieh, C.-M., Jan, F.-C. & Zhang, L. A simplified assessment of how tree allocation, wind environment, and shading affect human comfort. *Urban For. Urban Green.* **18**, 126–137 (2016).
180. Buccolieri, R., Santiago, J.-L., Rivas, E. & Sánchez, B. Reprint of: Review on urban tree modelling in CFD simulations: Aerodynamic, deposition and thermal effects. *Urban For. Urban Green.* **37**, 56–64 (2019).
181. Morakinyo, T. E., Ouyang, W., Lau, K. K.-L., Ren, C. & Ng, E. Right tree, right place (urban canyon): tree species selection approach for optimum urban heat mitigation — development and evaluation. *Sci. Total Environ.* **719**, 137461 (2020).
- Examines tree selection and placement for shade optimization.**
182. Jim, C. Y. & Chen, W. Y. Assessing the ecosystem service of air pollutant removal by urban trees in Guangzhou (China). *J. Environ. Manag.* **88**, 665–676 (2008).
183. Schwartz, A., Turbé, A., Simon, L. & Julliard, R. Enhancing urban biodiversity and its influence on city-dwellers: an experiment. *Biol. Conserv.* **171**, 82–90 (2014).
184. Twohig-Bennett, C. & Jones, A. The health benefits of the great outdoors: a systematic review and meta-analysis of greenspace exposure and health outcomes. *Environ. Res.* **166**, 628–637 (2018).
185. Han, Y., Taylor, J. E. & Pisello, A. L. Toward mitigating urban heat island effects: Investigating the thermal-energy impact of bio-inspired retro-reflective building envelopes in dense urban settings. *Energy Build.* **102**, 380–389 (2015).
186. Aram, F., Higueras Garcia, E., Solgi, E. & Mansournia, S. Urban green space cooling effect in cities. *Heliyon* **5**, e01339 (2019).
187. Wang, Y. & Akbari, H. The effects of street tree planting on urban heat island mitigation in Montreal. *Sustain. Cities Soc.* **27**, 122–128 (2016).

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